

# Spatio-temporal variations of soil water content and salinity around individual *Tamarix ramosissima* in a semi-arid saline region of the upper Yellow River, Northwest China

YANG Benman, WANG Ruoshui, XIAO Huijie\*, CAO Qiqi, LIU Tao

College of Soil and Water Conservation, Beijing Forestry University, Beijing 100083, China

**Abstract:** Ecological restoration by *Tamarix* plants on semi-arid saline lands affects the accumulation, distribution patterns and related mechanisms of soil water content and salinity. In this study, spatio-temporal variations of soil water content and salinity around natural individual *Tamarix ramosissima* Ledeb. were investigated in a semi-arid saline region of the upper Yellow River, Northwest China. Specifically, soil water content, electrical conductivity ( $EC_e$ ), sodium adsorption ratio ( $SAR_e$ ), and salt ions (including  $Na^+$ ,  $K^+$ ,  $Ca^{2+}$ ,  $Mg^{2+}$  and  $SO_4^{2-}$ ) were measured at different soil depths and at different distances from the trunk of *T. ramosissima* in May, July, and September 2016. The soil water content at the 20–80 cm depth was significantly lower in July and September than in May, indicating that *T. ramosissima* plants absorb a large amount of water through the roots during the growing period, leading to the decreasing of soil water content in the deep soil layer. At the 0–20 cm depth, there was a salt island effect around individual *T. ramosissima*, and the  $EC_e$  differed significantly inside and outside the canopy of *T. ramosissima* in May and July. Salt bioaccumulation and stemflow were two major contributing factors to this difference. The  $SAR_e$  at the 0–20 cm depth was significantly different inside and outside the canopy of *T. ramosissima* in the three sampling months. The values of  $SAR_e$  at the 60–80 cm depth in May and July were significantly higher than those at the 0–60 cm depth and higher than that at the corresponding depth in September. The distribution of  $Na^+$  in the soil was similar to that of the  $SAR_e$ , while the concentrations of  $K^+$ ,  $Ca^{2+}$ , and  $Mg^{2+}$  showed significant differences among the sampling months and soil depths. Both season and soil depth had highly significant effects on soil water content,  $EC_e$  and  $SAR_e$ , whereas distance from the trunk of *T. ramosissima* only significantly affected  $EC_e$ . Based on these results, we recommend co-planting of shallow-rooted salt-tolerant species near the *Tamarix* plants and avoiding planting herbaceous plants inside the canopy of *T. ramosissima* for afforestation in this semi-arid saline region. The results of this study may provide a reference for appropriate restoration in the semi-arid saline regions of the upper Yellow River.

**Keywords:** *Tamarix ramosissima*; soil water content; electrical conductivity; sodium adsorption ratio; saline soil; Yellow River

**Citation:** YANG Benman, WANG Ruoshui, XIAO Huijie, CAO Qiqi, LIU Tao. 2018. Spatio-temporal variations of soil water content and salinity around individual *Tamarix ramosissima* in a semi-arid saline region of the upper Yellow River, Northwest China. *Journal of Arid Land*, 10(1): 101–114. <https://doi.org/10.1007/s40333-017-0072-9>

\*Corresponding author: XIAO Huijie (E-mail: [herr\\_xiao@hotmail.com](mailto:herr_xiao@hotmail.com))

Received 2017-06-08; revised 2017-10-11; accepted 2017-10-24

© Xinjiang Institute of Ecology and Geography, Chinese Academy of Sciences, Science Press and Springer-Verlag GmbH Germany, part of Springer Nature 2018

## 1 Introduction

*Tamarix* (Tamaricaceae) plants are dominant tree species distributed in the semi-arid saline regions of the upper Yellow River. More than 10 species of *Tamarix* are distributed in these regions, and these species are typical salt-secreting halophytes (Sookbirsingh et al., 2010). In China, *Tamarix* plants are considered as excellent species for soil and water conservation as well as for afforestation in saline regions (Zhang et al., 2002). Large-scale planting of *Tamarix* plants can influence the distributions of soil water content and salinity in saline regions. The spatio-temporal variations of soil water content and salinity can, in turn, largely determine the spatial distributions of other vegetation types (Wu et al., 2009; He et al., 2014, 2015). These cascading effects are likely to influence how vegetation restoration occurs in saline regions. The semi-arid saline regions of the upper Yellow River are mostly located in the Yellow River irrigation area, where the spatio-temporal variations of soil water content and salinity are complex due to special hydrological conditions such as uneven rainfall distribution, shallow groundwater depth, and high salinity of groundwater. The spatio-temporal variations of soil water content and salinity around individual *Tamarix* plants are unclear in these regions. Such variations are closely related to regional vegetation restoration and vegetation allocation. Therefore, identifying the relationships between the *Tamarix* plants and spatio-temporal variations of soil water content and salinity in the semi-arid saline regions of the upper Yellow River will be important for vegetation restoration and vegetation allocation in future.

It has been shown that salt bioaccumulation could cause a salt island effect around individual *Tamarix* stands (Grubb et al., 1997), reflecting the spatial heterogeneity of soil on a small scale. The salt island effect also has a profound impact on the vegetation pattern of an ecosystem on a large scale (Hillerislambers et al., 2001). Some studies have observed a significant increase in soil salinity under the canopy of *Tamarix* plants in different land-cover types, including riparian corridors (Ladenburger et al., 2006), deserts (Yin et al., 2010), saline-alkali lands (Zhang and Chen, 2015) and coastal wetlands (Liu et al., 2017). However, Ohrtman et al. (2012) noted that soil salinity under the canopy of *Tamarix* plants was lower than that in an adjacent area without woody cover under non-flooded conditions. Additionally, Lesica and DeLuca (2004) and Li et al. (2007) found no difference in soil salinity inside and outside the canopy of *Tamarix* plants. It should be mentioned that previous related studies (e.g., Ladenburger et al., 2006; Yin et al., 2010; Zhang and Chen, 2015) mainly focused on the enrichment effect of *Tamarix* plants on soil salinity at a single time point. Numerous studies on soil water under *Tamarix* plants have also shown that the distribution of soil water varies greatly from region to region. Di Tomaso (1998) and Xia et al. (2016) noted that *Tamarix* plants could reduce soil water content. In contrast, Guan et al. (2009) found that soil water storage capacity was higher in the root zone of *Tamarix chinensis* Lour. than in an adjacent unplanted area in a saline coastal area of Shandong Province, China.

Investigating the spatio-temporal distributions of soil water content and salinity on the individual plant scale plays an important role in optimal vegetation management (Parkin, 1993). To date, few studies have investigated the spatio-temporal variations of soil water content and salinity around individual *Tamarix* plants in the semi-arid saline regions of the upper Yellow River. In the present study, we analyzed the spatial distributions and seasonal variations of soil water content, electrical conductivity ( $EC_e$ ), sodium adsorption ratio ( $SAR_e$ ), and salt ions at different soil depths and at different distances from the trunk of *Tamarix ramosissima* Ledeb. during the growing period in a semi-arid saline region of the upper Yellow River. Since *T. ramosissima* is a natural species commonly found in this study region, we hope that this study could provide a reference for the restoration and spatial allocation of vegetation in the semi-arid saline regions of the upper Yellow River.

## 2 Materials and methods

### 2.1 Study area

The study area was located on Shuxin Forest Farm of Xiaoba Town, Qingtongxia City, Ningxia

Hui Autonomous Region, Northwest China (38°01'N, 105°56'E; 1139 m a.s.l.). This region lies within the irrigation area of Yinchuan Plain along the upper reaches of the Yellow River. It should be noted that the region was previously a saline lake. The study area is characterized by a temperate continental climate with low precipitation (mean annual precipitation of 260.7 mm) and high evaporation (mean annual evaporation of 1296 mm). The annual mean temperature is 8.5°C and the mean annual sunshine hour is 2295 h. The average groundwater level is approximately 1.5 m and the total dissolved solid concentration of groundwater is in the range of 3–13 g/L. Native plant species include *T. ramosissima*, *Phragmitesaustralis* (Cav.) Trin. exStued., *Suaeda salsa* (L.) Pall., *Achnatherumsplendens* (Trin.) Nevski., and *Agropyroncristatum*(Linn.) Gaertn. The soil of the study area is sandy soil.

**2.2 Sampling methods**

**2.2.1 Characteristics of sampling plot**

A sampling plot with an area of 20 m×20 m was selected within the natural *T. ramosissima* community in the study area. Three living individual *T. ramosissima* plants in the sampling plot were selected as the targets for investigation. The selected plants had similar canopy size, height, and stem diameter measurements. Morphological traits of the selected *T. ramosissima* plants and characteristics of the understory vegetation are shown in Table 1. Measurements of soil water content and salinity were carried out on 10 May (spring), 15 July (summer) and 21 September (autumn), 2016. Table 1 shows the basic soil physical characteristics at the 0–80 cm soil depth in the sampling plot. The soil in the sampling plot is sandy soil. Furthermore, the soil is saline-sodic with high EC<sub>e</sub> and SAR<sub>e</sub>, and the dominant salts are sodium and sulfate.

**Table 1** Characteristics of the sampling plot

Morphological traits of the selected <i>T. ramosissima</i>			Understory vegetation			Soil texture				
Mean height (m)	Stem diameter (cm)	Canopy size (m×m)	Coverage (%)	Dominant species	Soil EC <sub>e</sub> (mS/cm)	Soil pH <sub>e</sub>	Soil SAR <sub>e</sub> (mmol/L) <sup>0.5</sup>	Sand (%)	Silt (%)	Clay (%)
3.99 (0.10)	5.02 (0.28)	4.19 (0.06)×4.25 (0.25)	<5% in May; 20%–30% in Jul; <10% in Sep	<i>Suaeda salsa</i> and <i>Phragmitesau</i> <i>stralis</i>	11.29 (0.56)	7.77 (0.08)	24.34 (3.86)	81.75 (2.96)	8.50 (2.09)	9.75 (1.27)

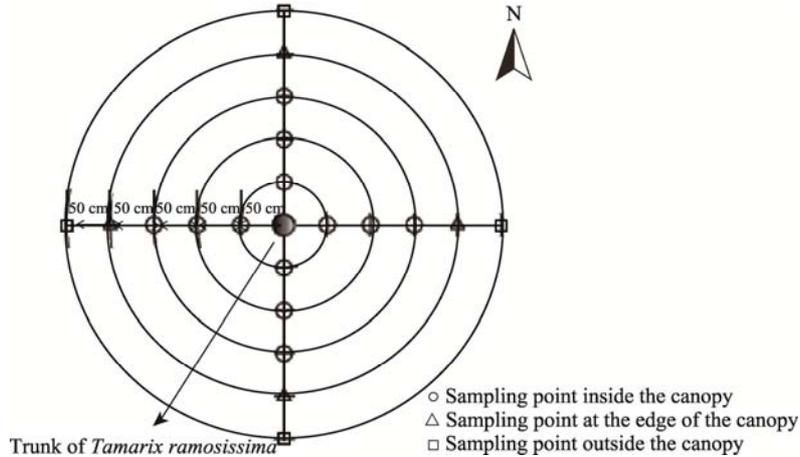
Note: EC<sub>e</sub>, electrical conductivity; SAR<sub>e</sub>, sodium adsorption ratio. Values in the parentheses represent standard deviations.

**2.2.2 Soil sampling**

Soil samples were collected using the concentric circle method (Zhang et al., 2016). For each individual *T. ramosissima*, we taken the trunk as the center of measurement and collected soil samples in four directions using a 5-cm-diameter soil auger. As shown in Figure 1, the sampling points were distributed at 50, 100, 150, 200 and 250 cm from the trunk in four directions. Specifically, three sampling points (50, 100, and 150 cm from the trunk) were situated inside the canopy of *T. ramosissima*, one sampling point (200 cm from the trunk) was located at the edge of the canopy, and one sampling point (250 cm from the trunk) was located outside the canopy. The soil was sampled at four depths (0–20, 20–40, 40–60, and 60–80 cm) with three replicates. Finally, a total of 720 soil samples (5 sampling points×4 soil depths×4 directions×3 repetitions×3 sampling months) were collected. For each soil sample, a part of the sample was separated out to measure soil water content and the remainder was stored in a sealed plastic bag to measure the other physical and chemical properties.

**2.3 Measurements and calculations**

Rainfall and air temperature data were collected using a small weather station (HOBOWare U30, Onset, USA) during the growing season of *T. ramosissima*. Soil water content was measured using the oven-drying method. Soil samples used for measuring the other physical and chemical properties were air-dried and then sieved with a 1-mm mesh sieve. Soil EC<sub>e</sub>, SAR<sub>e</sub>, pH<sub>e</sub> and concentrations of soluble ions (including Na<sup>+</sup>, K<sup>+</sup>, Ca<sup>2+</sup>, Mg<sup>2+</sup> and SO<sub>4</sub><sup>2-</sup>) were measured for all



**Fig. 1** Layout of the sampling points

samples as saturated paste extract. Soil texture was determined using a laser particle size analyzer (S5300, Microtrac, USA) after the soil was treated with dispersant. Soil  $\text{pH}_e$  and  $\text{EC}_e$  were determined using a pH meter (PHS-3C, Leici, China) and a conductivity meter (DDS-307, Leici, China), respectively. The concentrations of soluble ions were determined using an Inductively Coupled Plasma Optical Emission Spectrometer (Optima 5300DV, PerkinElmer, USA). Soil  $\text{SAR}_e$  was calculated using the following formula (Sou/Dakouré, 2013):

$$\text{SAR}_e = \frac{C_{\text{Na}^+}}{\sqrt{\frac{1}{2}(C_{\text{Ca}^{2+}} + C_{\text{Mg}^{2+}})}}, \quad (1)$$

where,  $C_{\text{Na}^+}$  is the concentration of  $\text{Na}^+$  (mmol/L);  $C_{\text{Ca}^{2+}}$  is the concentration of  $\text{Ca}^{2+}$  (mmol/L); and  $C_{\text{Mg}^{2+}}$  is the concentration of  $\text{Mg}^{2+}$  (mmol/L).

## 2.4 Data analysis

Soil water content and salinity data were analyzed using SPSS software version 18.0 (IBM, USA). One-way analysis of variance and least significant difference tests were used to analyze the differences in soil properties. In addition, Student-Newman-Keuls test was used to determine the effects of season, sampling location, soil depth and their interactions on soil water content and salinity. Differences of  $P < 0.05$  were considered significant. Furthermore, Microsoft Excel 2007 was used for data analysis and mapping.

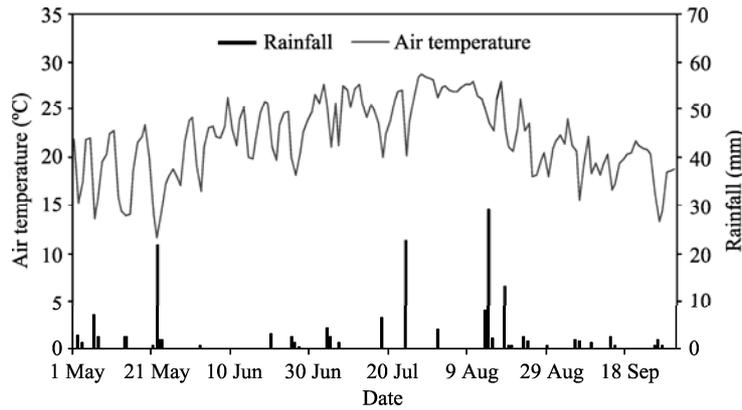
## 3 Results

### 3.1 Variations in rainfall and air temperature during the growing season of *T. ramosissima*

The total rainfall was 193.1 mm for the whole growing season of *T. ramosissima*, with 52.3% of total rainfall occurring during the high-growth period in July and August. As shown in Figure 2, rainfall in May 2016 (41.2 mm) was the highest among the three sampling months (i.e., May, July, and September), and it was also higher than the long-term average monthly rainfall in May (19.0 mm) in the region. Rainfall in September 2016 (11.5 mm) was the lowest among the three sampling months, and it was also lower than the long-term average monthly rainfall in September (26.0 mm). Furthermore, rainfall in July 2016 (37.8 mm) was slightly lower than the long-term average monthly rainfall in July (41.0 mm). The mean air temperature was highest in July (25.5°C), followed by September (19.4°C) and May (18.6°C). Air temperature showed irregular fluctuations during the whole growing season of *T. ramosissima* (Fig. 2).

### 3.2 Basic soil properties

Statistical analyses of soil water content,  $\text{EC}_e$ , and  $\text{SAR}_e$  at different soil depths around individual



**Fig. 2** Rainfall and air temperature during the growing season of *T. ramosissima*

*T. ramosissima* are shown in Table 2. Soil water content generally increased with increasing soil depth in all sampling months, with exceptions of 20–40 cm depth in July and September. Specifically, soil water content at the 20–40 cm depth was lower than that at the 0–20 cm depth in July and September. There were significant differences in the soil water content at different soil depths among the three sampling months (May, July, and September;  $P < 0.05$ ). The values of soil water content at the 60–80 cm depth were 4.9, 3.2, and 1.4 times higher than those at the 0–20 cm depth in May, July, and September, respectively. Table 2 shows that the CV (coefficient of variation) of soil water content at different soil depths in different sampling months varied from 31% to 80%. The CV of soil water content at the 0–20 cm depth was greater than that at the 20–80 cm depth in May and September. There were small differences in the CV of soil water content among different soil depths in July.

**Table 2** Statistical parameters of soil water content,  $EC_e$ , and  $SAR_e$  around individual *T. ramosissima*

Month	Soil depth (cm)	Soil water content			$EC_e$			$SAR_e$		
		Mean (%)	SD (%)	CV (%)	Mean (mS/cm)	SD (mS/cm)	CV (%)	Mean (mmol/L) <sup>0.5</sup>	SD (mmol/L) <sup>0.5</sup>	CV (%)
May	0–20	2.96 <sup>dAB</sup>	1.89	64.0	7.49 <sup>cB</sup>	6.10	81.4	14.90 <sup>bB</sup>	5.02	33.7
	20–40	6.05 <sup>cA</sup>	2.03	33.5	7.92 <sup>cB</sup>	4.50	56.8	15.62 <sup>bB</sup>	5.18	33.2
	40–60	9.18 <sup>bA</sup>	2.87	31.3	13.27 <sup>bA</sup>	6.65	50.1	20.67 <sup>bB</sup>	11.40	55.1
	60–80	14.46 <sup>aA</sup>	4.87	33.7	16.48 <sup>aA</sup>	6.39	38.8	51.79 <sup>aA</sup>	38.61	74.6
Jul	0–20	2.34 <sup>eB</sup>	1.56	66.7	11.73 <sup>bA</sup>	10.29	87.7	19.90 <sup>bA</sup>	6.92	34.8
	20–40	2.07 <sup>cC</sup>	0.99	48.1	9.97 <sup>bA</sup>	6.03	60.4	20.23 <sup>bA</sup>	7.03	34.8
	40–60	4.21 <sup>bB</sup>	2.86	67.9	15.02 <sup>aA</sup>	8.07	53.7	26.32 <sup>bA</sup>	14.86	56.5
	60–80	7.51 <sup>aB</sup>	4.48	59.7	16.53 <sup>aA</sup>	8.84	53.5	52.40 <sup>aA</sup>	45.47	86.8
Sep	0–20	3.48 <sup>bA</sup>	2.76	79.5	8.38 <sup>bcB</sup>	6.50	77.6	15.17 <sup>cB</sup>	5.25	34.6
	20–40	3.12 <sup>bB</sup>	0.97	31.1	7.50 <sup>cB</sup>	5.49	73.2	15.88 <sup>bcB</sup>	5.49	34.6
	40–60	4.27 <sup>aB</sup>	1.83	42.8	10.04 <sup>bB</sup>	6.17	61.5	17.87 <sup>bB</sup>	6.36	35.6
	60–80	4.97 <sup>aC</sup>	2.18	43.8	16.06 <sup>aA</sup>	8.21	51.1	21.27 <sup>ab</sup>	7.73	36.3

Note: SD, standard deviation; CV, coefficient of variation. Different lowercase letters indicate significant differences at different soil depths in the same sampling month ( $P < 0.05$ ); different capital letters indicate significant differences in different sampling months at the same soil depth ( $P < 0.05$ ).

The variation of  $EC_e$  in the vertical soil profile was consistent with that of soil water content. That is, the  $EC_e$  generally increased with increasing soil depth with exceptions of  $EC_e$  at the 20–40 cm depth in July and September. Also, the  $EC_e$  at the 20–40 cm depth was lower than that at the 0–20 cm depth in July and September. There were significant differences in the  $EC_e$  among different soil depths in May and September ( $P < 0.05$ ), and the values of  $EC_e$  at the 60–80 cm depth were 2.2 and 1.9 times higher than those at the 0–20 cm depth, respectively. Additionally, there was a significant difference of  $EC_e$  between the 0–40 and 40–80 cm depths in July ( $P < 0.05$ ), with

the latter being 31.2% higher than the former. The  $EC_e$  first increased and then decreased during the three sampling months at all soil depths. The CV of  $EC_e$  was highest at the 0–20 cm depth in all sampling months.

As similar as soil water content and  $EC_e$ , the  $SAR_e$  also increased with increasing soil depth. The  $SAR_e$  at the 60–80 cm depth showed distinct peaks in May and July, and was significantly higher than that at the 0–60 cm depth for the same months ( $P < 0.05$ ). The values of  $SAR_e$  at the 60–80 cm depth were 3.5 and 2.6 times higher than those at the 0–20 cm depth ( $P < 0.05$ ) in May and July, respectively. There were significant differences in the  $SAR_e$  among different soil depths in September ( $P < 0.05$ ), and the  $SAR_e$  at the 60–80 cm depth was 1.4 times higher than that at the 0–20 cm depth. Furthermore, the CV of  $SAR_e$  was highest at the 60–80 cm depth.

As shown in Table 3, the  $pH_e$  at the 60–80 cm depth was significantly higher than that at the 0–60 cm depth in May and July ( $P < 0.05$ ), whereas the  $pH_e$  at the 0–40 cm depth was significantly higher than that at the 40–80 cm depth in September ( $P < 0.05$ ). The  $pH_e$  value at each soil depth in May was significantly higher than those in July and September ( $P < 0.05$ ). The cations in the soil around individual *T. ramosissima* were dominated by  $Na^+$ , which was mainly distributed in the deep soil depths. The  $Na^+$  concentration at the 40–80 cm depth was significantly higher than that at the 0–40 cm depth in all sampling months ( $P < 0.05$ ). The  $Na^+$  concentration was found to be highest in July. The concentrations of  $K^+$ ,  $Ca^{2+}$ , and  $Mg^{2+}$  were relatively low, and all of them showed significant differences among the sampling months and soil depths ( $P < 0.05$ ). The concentration of  $SO_4^{2-}$  generally increased with increasing soil depth. There was no significant difference of  $SO_4^{2-}$  concentration at the 60–80 cm depth among the sampling months. The CV values of  $Na^+$ ,  $Ca^{2+}$ , and  $Mg^{2+}$  concentrations were highest at the 0–20 cm depth in all sampling months.

### 3.3 Variations in soil water content and salinity at different distances from individual *T. ramosissima*

#### 3.3.1 Variations in soil water content

Variations of soil water content at different soil depths and at different distances from the trunk of *T. ramosissima* showed different trends for the three sampling months. In May, the soil water content at the 0–20 cm depth gradually decreased with increasing distance from the trunk of *T. ramosissima*, and the soil water content inside the canopy (50 cm from the trunk) was two times higher than that outside the canopy (250 cm from the trunk) (Fig. 3a). At the 20–40 cm depth, the soil water content initially decreased and then increased with increasing distance from the trunk, and the difference was significant between the 50- and 150-cm distances from the trunk ( $P < 0.05$ ). At the 40–80 cm depth, the soil water content did not significantly differ inside and outside the canopy. In July, the highest soil water content at the 0–20 cm depth occurred outside the canopy (250 cm from the trunk), while the lowest value occurred at the edge of the canopy (200 cm from the trunk) (Fig. 3b). The soil water content inside the canopy increased with increasing distance from the trunk. There was no significant difference in the soil water content inside and outside the canopy at the 0–60 cm depth. However, at the 60–80 cm depth, the soil water content outside the canopy was significantly higher than those at the edge of canopy and inside the canopy ( $P < 0.05$ ). In September, variation of soil water content at the 0–20 cm depth (Fig. 3c) was similar to that observed in May (Fig. 3a). That is, in September, soil water content at the 0–20 cm depth decreased with increasing distance from the trunk (Fig. 3c). Furthermore, there was no significant difference in the soil water content inside and outside the canopy across the 0–80 cm depth.

#### 3.3.2 Variations in soil electrical conductivity ( $EC_e$ )

The spatio-temporal distributions of soil  $EC_e$  around individual *T. ramosissima* showed substantial differences at different distances from the trunk. As shown in Figure 4, at the 0–20 cm depth, the  $EC_e$  values at the 50- and 100-cm distances from the trunk were higher than those at the other sampling points in all sampling months, demonstrating an obvious salt island effect around the individual *T. ramosissima*. In May, the  $EC_e$  at the 0–20 cm depth was significantly higher inside the canopy (50 and 100 cm from the trunk) than outside the canopy (250 from the trunk) and at the edge of the canopy (200 from the trunk) ( $P < 0.05$ ; Fig. 4a). At the 20–40 cm depth, the

**Table 3** Statistical parameters of concentrations of soluble ions and  $\text{pH}_e$  values around individual *T. ramosissima*

Month	Soil depth (cm)	$\text{Na}^+$ concentration			$\text{Ca}^{2+}$ concentration			$\text{Mg}^{2+}$ concentration		
		Mean (g/kg)	SD (g/kg)	CV (%)	Mean (g/kg)	SD (g/kg)	CV (%)	Mean (g/kg)	SD (g/kg)	CV (%)
May	0–20	1.10 <sup>ab</sup>	0.84	76.4	0.26 <sup>ba</sup>	0.25	96.2	0.11 <sup>ba</sup>	0.12	109.1
	20–40	1.20 <sup>ab</sup>	0.69	57.5	0.28 <sup>ba</sup>	0.19	67.9	0.10 <sup>ba</sup>	0.09	90.0
	40–60	2.11 <sup>ba</sup>	1.20	56.9	0.45 <sup>aA</sup>	0.20	44.4	0.21 <sup>aA</sup>	0.15	71.4
	60–80	3.08 <sup>aA</sup>	1.28	41.6	0.21 <sup>bB</sup>	0.19	90.5	0.17 <sup>bB</sup>	0.14	82.4
Jul	0–20	1.74 <sup>ba</sup>	1.52	87.4	0.38 <sup>aA</sup>	0.41	107.9	0.16 <sup>aA</sup>	0.21	131.3
	20–40	1.55 <sup>ba</sup>	0.97	62.6	0.29 <sup>ba</sup>	0.21	72.4	0.11 <sup>ba</sup>	0.09	81.8
	40–60	2.49 <sup>aA</sup>	1.50	60.2	0.40 <sup>aAB</sup>	0.21	52.5	0.18 <sup>aA</sup>	0.12	66.7
	60–80	3.11 <sup>aA</sup>	1.92	61.7	0.24 <sup>bB</sup>	0.21	87.5	0.15 <sup>abB</sup>	0.12	80.0
Sep	0–20	1.15 <sup>ab</sup>	0.80	69.6	0.32 <sup>bcA</sup>	0.36	112.5	0.12 <sup>ba</sup>	0.13	108.3
	20–40	1.09 <sup>ab</sup>	0.75	68.8	0.26 <sup>ca</sup>	0.26	100.0	0.09 <sup>ba</sup>	0.09	100.0
	40–60	1.48 <sup>bb</sup>	0.93	62.8	0.35 <sup>bb</sup>	0.25	71.4	0.12 <sup>bb</sup>	0.10	83.3
	60–80	2.39 <sup>ab</sup>	1.25	52.3	0.53 <sup>aA</sup>	0.30	56.6	0.27 <sup>aA</sup>	0.21	77.8

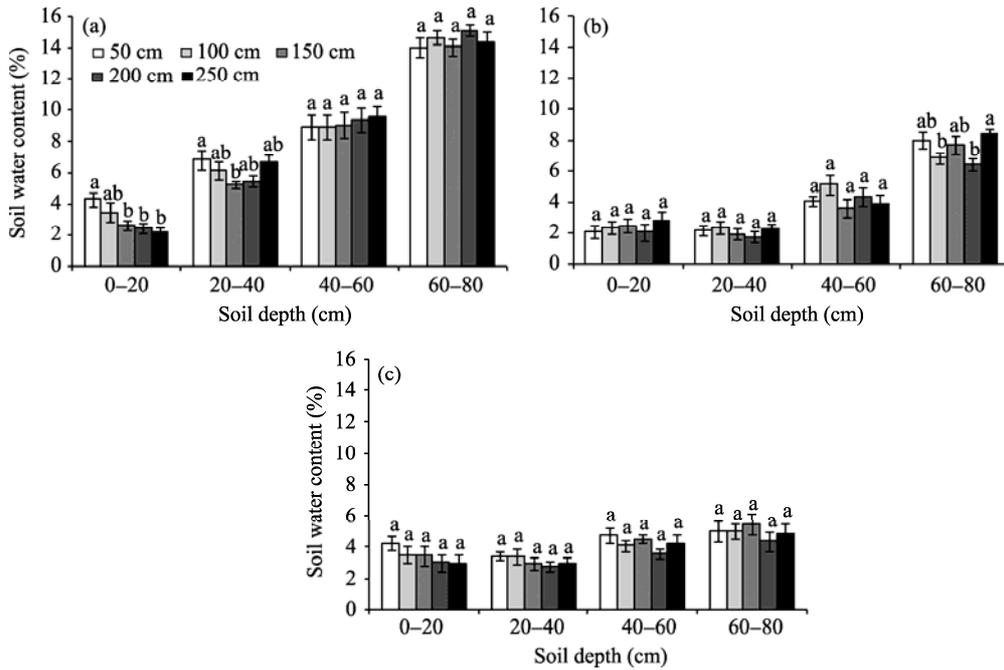
Month	Soil depth (cm)	$\text{K}^+$ concentration			$\text{SO}_4^{2-}$ concentration			$\text{pH}_e$		
		Mean (g/kg)	SD (g/kg)	CV (%)	Mean (g/kg)	SD (g/kg)	CV (%)	Mean	SD	CV (%)
May	0–20	0.09 <sup>ac</sup>	0.05	55.6	0.87 <sup>ba</sup>	0.63	72.4	7.76 <sup>ba</sup>	0.23	3.0
	20–40	0.05 <sup>bc</sup>	0.02	40.0	1.11 <sup>baB</sup>	0.77	69.4	7.58 <sup>ca</sup>	0.23	3.0
	40–60	0.04 <sup>bb</sup>	0.02	50.0	2.46 <sup>aA</sup>	1.43	58.1	7.57 <sup>ca</sup>	0.26	3.4
	60–80	0.05 <sup>bb</sup>	0.02	40.0	2.84 <sup>aA</sup>	1.64	57.7	8.18 <sup>aA</sup>	0.63	7.7
Jul	0–20	0.24 <sup>aA</sup>	0.22	91.7	1.17 <sup>ba</sup>	0.90	76.9	7.49 <sup>bb</sup>	0.22	2.9
	20–40	0.18 <sup>abA</sup>	0.28	155.6	1.22 <sup>ba</sup>	0.94	77.0	7.39 <sup>eb</sup>	0.18	2.4
	40–60	0.15 <sup>ba</sup>	0.18	120.0	2.45 <sup>aA</sup>	1.71	69.8	7.37 <sup>eb</sup>	0.15	2.0
	60–80	0.13 <sup>ba</sup>	0.14	107.7	2.62 <sup>aA</sup>	1.97	75.2	7.78 <sup>ab</sup>	0.66	8.5
Sep	0–20	0.14 <sup>ab</sup>	0.09	64.3	1.01 <sup>ca</sup>	0.86	85.1	7.47 <sup>ab</sup>	0.20	2.7
	20–40	0.09 <sup>bb</sup>	0.05	55.6	0.85 <sup>cb</sup>	0.73	85.9	7.46 <sup>ab</sup>	0.26	3.5
	40–60	0.07 <sup>bb</sup>	0.06	85.7	1.33 <sup>bb</sup>	1.02	76.7	7.36 <sup>bb</sup>	0.23	3.1
	60–80	0.07 <sup>bb</sup>	0.04	57.1	2.69 <sup>aA</sup>	1.62	60.2	7.28 <sup>bc</sup>	0.21	2.9

Note: Different lowercase letters indicate significant differences at different soil depths in the same month ( $P < 0.05$ ); different capital letters indicate significant differences in different months at the same soil depth ( $P < 0.05$ ).

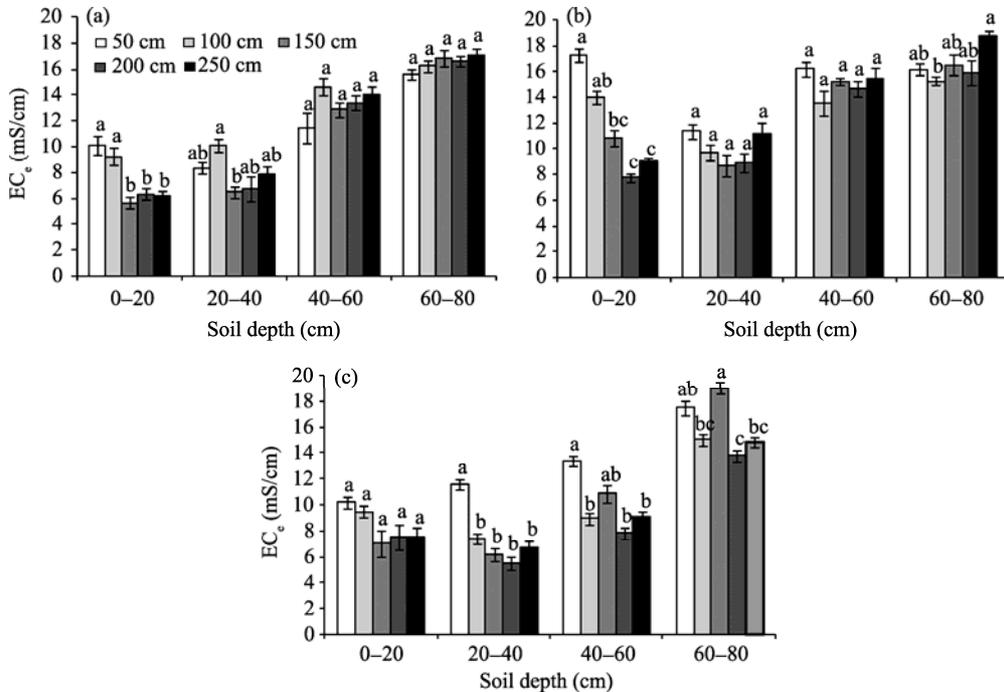
$\text{EC}_e$  inside the canopy was significantly higher at the 100-cm distance from the trunk than at the 150-cm distance from the trunk ( $P < 0.05$ ). At the 40–80 cm depth, no significant difference was found between the  $\text{EC}_e$  inside and outside the canopy. In July, at the 0–20 cm depth, the  $\text{EC}_e$  inside the canopy was 48.0% and 55.2% higher than those outside the canopy and at the edge of the canopy, respectively ( $P < 0.05$ ; Fig. 4b). At the 20–60 cm depth, there was no significant difference between the  $\text{EC}_e$  inside and outside the canopy. At the 60–80 cm depth, the  $\text{EC}_e$  outside the canopy was significantly higher than that inside the canopy at the 100-cm distance from the trunk ( $P < 0.05$ ). In September, the  $\text{EC}_e$  at the 0–20 cm depth did not significantly differ inside and outside the canopy, whereas the  $\text{EC}_e$  at the 20–80 cm depth was significantly different inside and outside the canopy ( $P < 0.05$ ; Fig. 4c).

### 3.3.3 Variations in soil sodium adsorption ratio ( $\text{SAR}_e$ )

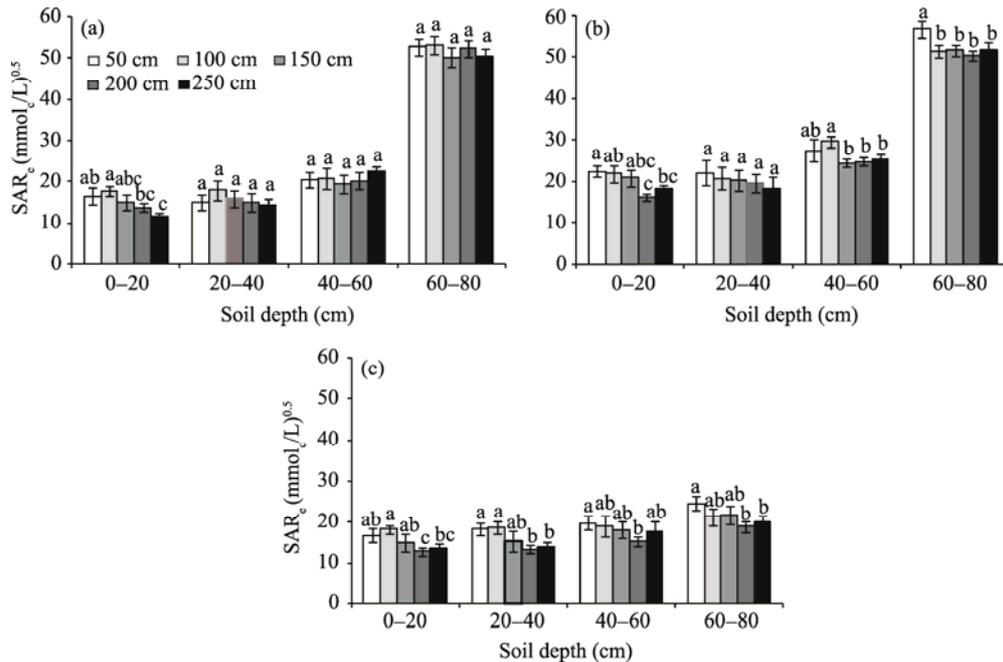
Similar to  $\text{EC}_e$ , the  $\text{SAR}_e$  at the 0–20 cm depth was significant high inside the canopy of *T. ramosissima* in all sampling months (Fig. 5). Specifically, the  $\text{SAR}_e$  value at the 0–20 cm depth was significantly higher inside the canopy at the 100-cm distance from the trunk than those outside the canopy (250 cm from the trunk) and at the edge of the canopy (200 cm from the trunk) ( $P < 0.05$ ) in May and September. Additionally, in July, the  $\text{SAR}_e$  at the 0–20 cm depth inside the canopy at the 50-cm distance from the trunk was 27.5% and 19.3% higher than those outside the canopy and at the edge of the canopy, respectively ( $P < 0.05$ ). In May, the  $\text{SAR}_e$  at the 20–80 cm



**Fig. 3** Variations of soil water content at different soil depths and at different distances from the trunk of *T. ramosissima* in May (a), July (b) and September (c). The legends indicate distances from the trunk of *T. ramosissima*. Different lowercase letters indicate significant difference among different distances from the trunk of *T. ramosissima* at the same soil depth ( $P < 0.05$ ). Bars mean standard deviations.



**Fig. 4** Variations of electrical conductivity (EC<sub>e</sub>) at different soil depths and at different distances from the trunk of *T. ramosissima* in May (a), July (b) and September (c). The legends indicate distances from the trunk of *T. ramosissima*. Different lowercase letters indicate significant differences among different distances from the trunk of *T. ramosissima* at the same soil depth ( $P < 0.05$ ). Bars mean standard deviations.



**Fig. 5** Variations of sodium adsorption ratio (SAR<sub>e</sub>) at different soil depths and at different distances from the trunk of *T. ramosissima* in May (a), July (b) and September (c). The legends indicate distances from the trunk of *T. ramosissima*. Different lowercase letters indicate significant differences among different distances from the trunk at the same soil depth ( $P < 0.05$ ). Bars mean standard deviations.

depth did not significantly differ inside and outside the canopy (Fig. 5a). In July, the SAR<sub>e</sub> at the 20–40 cm depth decreased with increasing distance from the trunk of *T. ramosissima*, and the SAR<sub>e</sub> inside the canopy at the 50-cm distance from the trunk was 16.6% higher than that outside the canopy (Fig. 5b). Moreover, there was a significant difference of SAR<sub>e</sub> at the 40–80 cm depth inside and outside the canopy ( $P < 0.05$ ). In September, the SAR<sub>e</sub> values at the 20–40 and 60–80 cm depths inside the canopy were significantly higher than those outside the canopy and at the edge of the canopy ( $P < 0.05$ ; Fig. 5c). The SAR<sub>e</sub> at the 40–60 cm depth inside the canopy was also significantly higher than that at the edge of the canopy ( $P < 0.05$ ). The SAR<sub>e</sub> at the 0–20 cm depth differed significantly inside and outside the canopy in all sampling months ( $P < 0.05$ ).

### 3.4 Effects of season, distance and soil depth on soil water content and salinity around individual *T. ramosissima*

Multivariate analysis of variance shows that season and soil depth had highly significant effects on soil water content, EC<sub>e</sub>, and SAR<sub>e</sub> around individual *T. ramosissima* ( $P < 0.01$ ; Table 4). However, distance from the trunk only significantly affected EC<sub>e</sub> around individual *T. ramosissima* ( $P < 0.05$ ), indicating that there were significant differences in EC<sub>e</sub> inside and outside the canopy of *T. ramosissima*, and that salt enrichment led to the formation of a salt island under the canopy. The relationship between season and soil depth was highly significant for soil water content and SAR<sub>e</sub> ( $P < 0.01$ ). However, the relationships between season and distance, and between distance and soil depth, and the relationship among these three factors were not significant either for soil water content or for soil salinity.

## 4 Discussion

### 4.1 Spatio-temporal variations of soil water content around individual *T. ramosissima*

Spatial heterogeneity of soil physical-chemical properties is common in arid and semi-arid regions, and plants (e.g., shrubs) contribute mostly to this variation (Zhang et al., 2001). In the horizontal direction, we found that soil water content in the surface soil layer (0–20 cm) increased

**Table 4** Multivariate analysis of variance for the effects of season, distance and soil depth on soil water content and salinity around individual *T. ramosissima*

Factors	F value		
	Soil water content	EC <sub>e</sub>	SAR <sub>e</sub>
Season	182.505**	9.963**	24.745**
Distance	1.014 <sup>ns</sup>	3.054*	0.923 <sup>ns</sup>
Soil depth	172.708**	46.599**	67.882**
Season×Distance	0.390 <sup>ns</sup>	0.896 <sup>ns</sup>	0.087 <sup>ns</sup>
Season×Soil depth	35.997**	1.869 <sup>ns</sup>	10.503**
Distance×Soil depth	0.266 <sup>ns</sup>	1.064 <sup>ns</sup>	0.085 <sup>ns</sup>
Season×Distance×Soil depth	0.483 <sup>ns</sup>	0.413 <sup>ns</sup>	0.041 <sup>ns</sup>

Note: \*, significant difference at  $P < 0.05$  level; \*\*, significant difference at  $P < 0.01$  level; <sup>ns</sup>, not significant.

with increasing distance from the trunk of *T. ramosissima* only in July, while the opposite trend was observed in May and September (Fig. 3). Additionally, soil water content in the surface soil layer (0–20 cm) inside the canopy of *T. ramosissima* was significantly higher than that at the edge of the canopy and outside the canopy of *T. ramosissima* in May. However, no consistent trends were found in soil water content at the 20–80 cm depth. Our results differ from the results of Yin et al. (2007), who found that soil water content inside the canopy of *Tamarix* shrubs was lower than that outside the canopy. In this study, the lower soil water content inside the canopy of *T. ramosissima* in July was perhaps induced by the transpiration through water uptake by the roots (Ehleringer et al., 1991; Mounisif et al., 2002). This inference is supported by the decreasing soil water content at the 20–60 cm depth over the study period.

In the vertical profile, the soil water content around individual *T. ramosissima* increased with increasing soil depth, and the CV of soil water content showed moderate variation at all depths (Table 2) (Nielsen and Bouma, 1985). The soil water content at the 20–80 cm depth in May was significantly higher than those in July and September, indicating that there was a clear seasonal variation of soil water content around individual *T. ramosissima*. This variation may be resulted from the interaction of climate, soil, and vegetation-related factors. The soil in the study area was sandy soil with poor water-holding capacity and high water permeability, thus soil water content is higher in the deep soil layer than in the surface soil layer. It should be noted that there was relatively high rainfall, low temperature, and weak evaporation before the sampling in May. In contrast, there was relatively low rainfall, high temperature, and strong evaporation before the sampling in July and September (Fig. 2). This may be the major reason that soil water content was overall higher in May than in July and September.

*Tamarix* plants can absorb water from multiple sources and they have the capacity to tolerate drought (Nippert et al., 2010; Yu et al., 2013; Wu et al., 2016). Chen et al. (2016) stated that groundwater and deep soil water are the main sources of water for *T. ramosissima*, and that the hydraulic lift by the roots ensures the shallow soil layer maintaining a certain level of soil water content. Zhou et al. (2013) also found that the absorbing roots of *T. ramosissima* are distributed in the deep soil layer, and deep soil water and groundwater contribute 90% of the water sources for *T. ramosissima*. Indeed, *Tamarix* is a highly water-consuming genera (Gay and Fritschen, 1979; Sala et al., 1996; Cleverly et al., 1997; Nagler et al., 2005). Schofield (1992) found that *Tamarix* plants can significantly lower the groundwater level via high water consumption. The water consumption of *Tamarix* plants is highest in summer, followed by autumn. In addition to groundwater, *Tamarix* plants also need to absorb water from the deep soil layer to maintain the normal growth both in summer and autumn. Consequently, the soil water content in the deep soil layer around *T. ramosissima* significantly decreased over the seasons. Soil water content in the surface soil layer was also likely influenced by the understory vegetation of *T. ramosissima*. Because the herbaceous vegetation around *T. ramosissima* had not germinated and the soil surface was exposed in May, the soil water content in the surface soil layer was slightly lower in May than in September. In July, evaporation was at its peak and vigorous herbaceous vegetation also increased water consumption. Thus, soil water content in the surface soil layer was lowest in July. To sum up, *T. ramosissima* consume a large amount of water during the growing season, resulting

in a low soil water content in the deep soil layer.

#### 4.2 Spatio-temporal variations of soil $EC_e$ around individual *T. ramosissima*

This study shows that the  $EC_e$  in the surface soil layer (0–20 cm) around individual *T. ramosissima* was higher inside the canopy than outside the canopy and at the edge of the canopy in all sampling months (Fig. 4). A salt island was formed around the *T. ramosissima* due to salt enrichment, which is consistent with the results of Brotherson and Field (1987) and Su et al. (2012). This phenomenon may be attributable to the following reasons. First, *Tamarix* plants are salt-secreting species that can absorb soil salt via the roots, transport the salt to the branches and leaves, and finally excrete it to the outside through litterfall, thereby returning the salt to the surface soil (Arndt et al., 2004). Moreover, the multi-stemmed hemispheroidal crowns of *Tamarix* plants are favorable for capturing and holding the litterfall (Bochet et al., 1999), thus maximizing the return of salt to the surface soil. Second, the stemflow can lead to a difference in salt distribution around plants. Návar (1993) showed that the rainwater collected by stemflow was more than five-fold higher than that intercepted by the canopy in three types of sub-shrubs in a semi-arid region. The *Tamarix* plants excrete absorbed salt to the outside through their salt glands. Salt excreted by the branches and leaves is leached from the canopy through rainwater and then intercepted by the litterfall under the canopy, thereby forming a high salinity area around the plants. In our study, the annual rainfall is relatively low and the rainfall amount in both July and September was lower than the respective average monthly rainfall amount. Thus, the stemflow of *T. ramosissima* was relatively low and the salt was likely concentrated (Fan and Hong, 2001), thereby increasing the enrichment of salt under the canopy. Third, the non-biological environment has an impact on salt distribution. The  $EC_e$  in the surface soil layer (0–20 cm) around individual *T. ramosissima* was significantly higher in July than in May and September, which differed from the results of Zhao et al. (2016), who found that the EC in the surface soil layer (0–20 cm) was highest in spring. This difference may be due to the fact that the stronger groundwater evaporation in the hot and dry summer led to the surface accumulation of salt in the current study. Thus, we surmise that our findings on the  $EC_e$  in the surface soil layer (0–20 cm) resulted from the interactions of salt bioaccumulation and stemflow of *T. ramosissima* and non-biological environment in the study area.

Generally speaking, evaporation is lower inside the canopy than outside the canopy because of the shading effect of the canopy. However, in this study, we found that in the surface soil layer (0–20 cm), the  $EC_e$  was significantly higher and the soil water content was lower inside the canopy than outside the canopy in July. This probably occurred because the *T. ramosissima* formed a high-salinity area around its main root zone, which inhibited the growth of understory herbaceous vegetation. The surface soil around the main root zone of *T. ramosissima* was therefore exposed and only covered with litterfall. Consequently, evaporation in summer was greater in the soil exposed under the canopy than that covered with vegetation outside the canopy.

In this study, the understory herbaceous plants around the canopy of *T. ramosissima* was dominated by *S. salsa*, which is a deep-rooted, salt-tolerant species (Yi and Wang, 2011). Absorption of salt by *S. salsa* might explain the presence of a low-salinity area at the 40 cm soil depth in July (Fig. 4b). *Tamarix* is highly salt-tolerant, with higher salt accumulation in the rhizosphere (Li et al., 2007). In our study, the salt around *T. ramosissima* mainly accumulated in the deep soil layer (Table 2), which is different from the results of Yin et al. (2010). The distribution characteristics of soil salinity around *T. ramosissima* in this study suggest that it is appropriate to plant shallow-rooted, salt-tolerant plants near the *Tamarix* plants, and that herbaceous species should not be planted inside the canopy of *T. ramosissima*. Furthermore, timely clearing of litterfall under the canopy of *Tamarix* plants can reduce the accumulation of salt in the surface soil layer and prevent the return of salt into the deep soil layer.

#### 4.3 Spatio-temporal variations of soil $SAR_e$ and salt ions around individual *T. ramosissima*

Similar to  $EC_e$ , the  $SAR_e$  in the surface soil layer (0–20 cm) around individual *T. ramosissima* was significantly high in all sampling months (Fig. 5). This indicates that salt enrichment under

the canopy of *T. ramosissima* was dominated by sodium, which agrees with the results of Zhang et al. (2016). The distribution characteristics of SAR<sub>e</sub> are associated with the selective absorption of salt ions by *Tamarix* plants. The *Tamarix* plants have salt glands that could selectively secrete salt ions (Shuyskaya et al., 2017). The secretion of Na<sup>+</sup> is markedly higher than those of Ca<sup>2+</sup> and Mg<sup>2+</sup>, which may be due to the high Na<sup>+</sup> concentration in leaves and branches of *Tamarix* plants (Glenn et al., 2012; Imada et al., 2013). In the vertical soil profile, we found that the SAR<sub>e</sub> was significantly higher at the 60–80 cm depth than at the 0–60 cm depth in May and July, and the SAR<sub>e</sub> at the 60–80 cm depth in September was significantly lower than those in May and July (Table 2; Fig. 5), which is likely related to the variations of soil water content and EC<sub>e</sub>. Li and Sun (2015) have suggested that the *Tamarix* plants can improve the saline environments by reducing the groundwater level. The deep roots of *Tamarix* plants absorb groundwater to replenish water lost by transpiration, resulting in a gradual drop in the groundwater level and an accumulation of salt in the deep soil layer (Li and Sun, 2015). During the growing period of *T. ramosissima*, a large number of Na<sup>+</sup> ions in the deep soil layer enters the plant via root absorption and then returns to the surface soil layer through secretion and litterfall (Imada et al., 2013). Moreover, the Na<sup>+</sup> ions may transfer to the deep soil layer along with the drop in the groundwater level. In this study, the significant decrease in Na<sup>+</sup> concentration and increases in Ca<sup>2+</sup> and Mg<sup>2+</sup> concentrations led to the SAR<sub>e</sub> at the 60–80 cm depth in September being significantly lower than those in May and July.

In the study area, the cations in the soil were mainly dominated by Na<sup>+</sup> and the enrichment effect was relatively clear for Na<sup>+</sup>. The Na<sup>+</sup> concentration at the 40–80 cm depth was significantly higher in May and July than in September. However, SO<sub>4</sub><sup>2-</sup>, which was relatively abundant in the study area, showed no significant difference at the 60–80 cm depth among different sampling months (Table 3). This also suggests that *T. ramosissima* mainly absorb Na<sup>+</sup>. The concentrations of Na<sup>+</sup>, Ca<sup>2+</sup>, and Mg<sup>2+</sup> all showed the largest variations in the surface soil layer (0–20 cm) in all sampling months. Thus, the SAR<sub>e</sub> in the surface soil layer showed significant differences inside and outside the canopy. Salt bioaccumulation was mainly associated with the absorption and transportation of salt by plant roots. Although the salt island effect was mainly present in the surface soil layer under the canopy, the rhizosphere activity also probably played an important role (Yin et al., 2012). The distribution characteristics of SAR<sub>e</sub> and salt ions around *T. ramosissima* indicate that individual *T. ramosissima* can enrich the Na<sup>+</sup> ions around the roots. Furthermore, *Tamarix* plants have the ability to improve the saline-alkali soil and they can play a better role when they are planted in sodium-type saline soils.

## 5 Conclusions

Spatio-temporal variations of soil water content and salinity around natural individual *T. ramosissima* were investigated in a semi-arid saline region of the upper Yellow River. The results show that in the semi-arid saline regions of the upper Yellow River, *T. ramosissima* continuously absorbed deep soil water via their roots during the growing period, resulting in the decreasing in soil water content in the deep soil layer. Soil EC<sub>e</sub> and SAR<sub>e</sub> at the 0–20 cm depth inside the canopy of *T. ramosissima* were significantly higher than those outside the canopy, exhibiting an obvious salt island effect. The accumulation of salt under the canopy of *T. ramosissima* was dominated by Na<sup>+</sup>. Both season and soil depth had highly significant effects on soil water content, EC<sub>e</sub>, and SAR<sub>e</sub>, while distance from the trunk of *T. ramosissima* had a significant effect only on EC<sub>e</sub>. Therefore, we recommend co-planting shallow-rooted salt-tolerant species near the *Tamarix* plants but avoiding planting herbaceous species inside the canopy of *Tamarix* plants. We further recommend choosing locations with a suitable groundwater level for planting *Tamarix* plants in the semi-arid saline regions of the upper Yellow River.

## Acknowledgments

This study was funded by the Fundamental Research Funds for the Central Universities (2016ZCQ06), the

Forestry Industry Research Special Funds for Public Welfare Projects (201504402) and the Application Technology of Seaweed Fertilizer Based on Desertification Control and Saline-alkali Soil Improvement (2016HXFWSBXY002). We thank Ningxia University and Shuxin Forest Farm of Xiaoba Town, Ningxia Hui Autonomous Region, China for field work and laboratory assistances.

## References

- Arndt S K, Arampatsis C, Foetzki A, et al. 2004. Contrasting patterns of leaf solute accumulation and salt adaptation in four Phreatophytic desert plants in a hyperarid desert with saline groundwater. *Journal of Arid Environments*, 59(2): 259–270.
- Bochet E, Rubio J L, Poesen J. 1999. Modified topsoil islands within patchy Mediterranean vegetation in SE Spain. *CATENA*, 38(1): 23–44.
- Brotherson J D, Field D. 1987. *Tamarix*: impacts of a successful weed. *Rangelands*, 9(3): 110–112.
- Chen Y P, Chen Y N, Xu C C, et al. 2016. The effects of groundwater depth on water uptake of *Populus euphratica* and *Tamarix ramosissima* in the hyperarid region of Northwestern China. *Environmental Science and Pollution Research*, 23(17): 17404–17412.
- Cleverly J R, Smith S D, Sala A, et al. 1997. Invasive capacity of *Tamarix ramosissima* in a Mojave Desert floodplain: the role of drought. *Oecologia*, 111(1): 12–18.
- Di Tomaso J M. 1998. Impact, biology, and ecology of saltcedar (*Tamarix* spp.) in the southwestern United States. *Weed Technology*, 12(2): 326–336.
- Ehleringer J R, Phillips S L, Schuster W S F, et al. 1991. Differential utilization of summer rains by desert plants. *Oecologia*, 88(3): 430–434.
- Fan H B, Hong W. 2001. Estimation of dry deposition and canopy exchange in Chinese fir plantations. *Forest Ecology and Management*, 147(2–3): 99–107.
- Gay L W, Fritschen L J. 1979. An energy budget analysis of water use by saltcedar. *Water Resources Research*, 15(6): 1589–1592.
- Glenn E P, Morino K, Nagler P L, et al. 2012. Roles of saltcedar (*Tamarix* spp.) and capillary rise in salinizing a non-flooding terrace on a flow-regulated desert river. *Journal of Arid Environments*, 79: 56–65.
- Grubb R T, Sheley R L, Carlstrom R D. 1997. Saltcedar (tamarisk). Montana State University Extension Service MT9710. Bozeman, USA: Montana State University.
- Guan H B, Wang X L, Ju D. 2009. Soiled modification and application of *Tamarix chinensis* on the saline soil. *Resource Development & Market*, 25(10): 918–921. (in Chinese)
- He B, Cai Y L, Ran W R, et al. 2014. Spatial and seasonal variations of soil salinity following vegetation restoration in coastal saline land in eastern China. *CATENA*, 118: 147–153.
- He B, Cai Y L, Ran W R, et al. 2015. Spatiotemporal heterogeneity of soil salinity after the establishment of vegetation on a coastal saline field. *CATENA*, 127: 129–134.
- Hillerislambers R, Rietkerk M, van den Bosch F, et al. 2001. Vegetation pattern formation in semi-arid grazing systems. *Ecology*, 82(1): 50–61.
- Imada S, Acharya K, Li Y P, et al. 2013. Salt dynamics in *Tamarix ramosissima* in the lower Virgin River floodplain, Nevada. *Trees*, 27(4): 949–958.
- Ladenburger C G, Hild A L, Kazmer D J, et al. 2006. Soil salinity patterns in *Tamarix* invasions in the Bighorn Basin, Wyoming, USA. *Journal of Arid Environments*, 65(1): 111–128.
- Lesica P, DeLuca T H. 2004. Is tamarisk allelopathic? *Plant and Soil*, 267(1–2): 357–365.
- Li J, Zhao C, Zhu H, et al. 2007. Effect of plant species on shrub fertile island at an oasis-desert ecotone in the South Junggar Basin, China. *Journal of Arid Environments*, 71(4): 350–361.
- Li Z J, Sun W Y. 2015. Saline-Alkali Soil Agro-Ecological Engineering. Beijing: Science Press, 93–97. (in Chinese)
- Liu J T, Rong Q Q, Zhao Y Y. 2017. Variations in soil nutrients and salinity caused by tamarisk in the coastal wetland of the Laizhou Bay, China. *Ecosphere*, 8(2): e01672, doi: 10.1002/ecs2.1672.
- Mounsif M, Wan C G, Sosebee R E. 2002. Effects of top-soil drying on saltcedar photosynthesis and stomatal conductance. *Journal of Range Management*, 55(1): 88–93.
- Nagler P L, Scott R L, Westenburg C, et al. 2005. Evapotranspiration on Western U.S. rivers estimated using the Enhanced Vegetation Index from MODIS and data from eddy covariance and Bowen ratio flux towers. *Remote Sensing of Environment*, 97(3): 337–351.
- Návar J. 1993. The causes of stemflow variation in three semi-arid growing species of northeastern Mexico. *Journal of*

- Hydrology, 145(1–2): 175–190.
- Nielsen D R, Bouma J. 1985. Soil Spatial Variability. Wageningen: PUDOC Scientific Publishers, 2–30.
- Nippert J B, Butler J J Jr, Kluitenberg G J, et al. 2010. Patterns of *Tamarix* water use during a record drought. *Oecologia*, 162(2): 283–292.
- Ohrtmann M K, Sher A A, Lair K D. 2012. Quantifying soil salinity in areas invaded by *Tamarix* spp. *Journal of Arid Environments*, 85: 114–121.
- Parkin T B. 1993. Spatial variability of microbial processes in soil—a review. *Journal of Environment Quality*, 22(3): 409–417.
- Sala A, Smith S D, Devitt D A. 1996. Water use by *Tamarix ramosissima* and associated phreatophytes in a Mojave desert floodplain. *Ecological Applications*, 6(3): 888–898.
- Schofield N J. 1992. Tree planting for dryland salinity control in Australia. *Agroforestry Systems*, 20(1–2): 1–23.
- Shuyskaya E V, Rakhmuklova Z F, Lebedeva M P, et al. 2017. Different mechanisms of ion homeostasis are dominant in the recretohalophyte *Tamarix ramosissima* under different soil salinity. *Acta Physiologiae Plantarum*, 39(3): 81, doi: 10.1007/s11738-017-2379-8.
- Sookbirsingh R, Castillo K, Gill T E, et al. 2010. Salt separation processes in the saltcedar *Tamarix ramosissima* (Ledeb.). *Communications in Soil Science and Plant Analysis*, 41(10): 1271–1281.
- Sou/Dakouré M Y, Mermoud A, Yacouba H, et al. 2013. Impacts of irrigation with industrial treated wastewater on soil properties. *Geoderma*, 200–201: 31–39.
- Su Y Z, Wang X F, Yang R, et al. 2012. Soil fertility, salinity and nematode diversity influenced by *Tamarix ramosissima* in different habitats in an arid desert oasis. *Environmental Management*, 50(2): 226–236.
- Wu G L, Jiang S W, Liu W Y, et al. 2016. Competition between *Populus euphratica* and *Tamarix ramosissima* seedlings under simulated high groundwater availability. *Journal of Arid Land*, 8(2): 157–171.
- Wu Y Y, Liu R C, Zhao Y G, et al. 2009. Spatial and seasonal variation of salt ions under the influence of halophytes, in a coastal flat in eastern China. *Environmental Geology*, 57(7): 1501–1508.
- Xia J B, Zhang S Y, Zhao X M, et al. 2016. Effects of different groundwater depths on the distribution characteristics of soil-*Tamarix* water contents and salinity under saline mineralization conditions. *CATENA*, 142: 166–176.
- Yi L P, Wang Z W. 2011. Root system characters in growth and distribution among three littoral halophytes. *Acta Ecologica Sinica*, 31(5): 1195–1202. (in Chinese)
- Yin C H, Feng G, Tian C Y, et al. 2007. Influence of tamarisk shrub on the distribution of soil salinity and moisture on the edge of Taklamakan Desert. *China Environmental Science*, 27(5): 670–675. (in Chinese)
- Yin C H, Feng G, Zhang F S, et al. 2010. Enrichment of soil fertility and salinity by tamarisk in saline soils on the northern edge of the Taklamakan Desert. *Agricultural Water Management*, 97(12): 1978–1986.
- Yin C H, Dong J Z, Shi Q M, et al. 2012. Salt island effect of halophytic shrubs in different habitats and its ecological implication. *Acta Pedologica Sinica*, 49(2): 289–295. (in Chinese)
- Yu T F, Feng Q, Si J H, et al. 2013. Patterns, magnitude, and controlling factors of hydraulic redistribution of soil water by *Tamarix ramosissima* roots. *Journal of Arid Land*, 5(3): 396–407.
- Zhang D Y, Yin L K, Pan B R. 2002. Biological and ecological characteristics of *Tamarix* L. and its effect on the ecological environment. *Science in China Series D: Earth Sciences*, 45(Suppl.): 18–22.
- Zhang H, Shi P J, Zheng Q H. 2001. Research progress in relationship between shrub invasion and soil heterogeneity in a natural semi-arid grassland. *Acta Phytocological Sinica*, 25(3): 366–370. (in Chinese)
- Zhang L H, Chen X B. 2015. Characteristics of ‘salt island’ and ‘fertile island’ for *Tamarix chinensis* and soil carbon, nitrogen and phosphorus ecological stoichiometry in saline-alkali land. *Chinese Journal of Applied Ecology*, 26(3): 653–658. (in Chinese)
- Zhang L H, Chen P H, Li J, et al. 2016. Distribution of soil salt ions around *Tamarix chinensis* individuals in the Yellow River Delta. *Acta Ecologica Sinica*, 36(18): 5741–5749. (in Chinese)
- Zhao Q Q, Bai J H, Liu Q, et al. 2016. Spatial and seasonal variations of soil carbon and nitrogen content and stock in a tidal salt marsh with *Tamarix chinensis*, China. *Wetlands*, 36(Suppl.): 145–152.
- Zhou H, Zheng X J, Tang L S, et al. 2013. Differences and similarities between water sources of *Tamarix ramosissima*, *Nitraria sibirica* and *Reaumuria soongorica* in the southeastern Junggar Basin. *Chinese Journal of Plant Ecology*, 37(7): 665–673. (in Chinese)