



Challenges for the sustainable use of water and land resources under a changing climate and increasing salinization in the Jizzakh irrigation zone of Uzbekistan

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Abstract: Jizzakh Province in Uzbekistan is one of the largest irrigated areas in Central Asia without natural drainage. In combination with aridity, climate change and extensive irrigation practices, this has led to the widespread salinization of agricultural land. The aim of this study was to identify opportunities to improve the reclamation status of the irrigated area and how best to effectively use the water resources in Jizzakh Province based on investigations conducted between 1995 and 2016. A database of field measurements of groundwater levels, mineralization and soil salinity conducted by the provincial Hydro-Geological Reclamation Expeditions was used in the study. The total groundwater mineralization was determined using a portable electric conductometer (Progress 1T) and the chloride concentration was determined using the Mohr method. The soil salinity analyses were conducted by applying two different methods: (1) the extraction and assessment of the soluble salt content, and (2) using an SM-138 conductivity sensor applied to a 1:1 mixture of soil sample and water. The analyses of the monitoring results and the salt balance in the "irrigation water–soil–drainage water" system clearly demonstrated that the condition of the irrigated land in the province was not significantly improved. Under these conditions, the stability of crop yields is achieved mainly through the use of large volumes of fertilizer. However, excess amounts of mineral fertilizers can also cause the salinization of soils. The average groundwater salinization value in most of the irrigated land (75.3%) fluctuated between 1.1 and 5.0 g/L, while the values were less than 1.0 g/L in 13.1% of the land and in the range of 5.1–10.0 g/L in 10.5% of the land. During the period of 1995–2016 the salinization level of the irrigated land in Jizzakh Province increased slightly and the area could be divided into the following classes: no salinity (17.7% of the total area), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%). Detailed studies of the salt balance in irrigated land, the impact of climate change, increased fertilizer use, and repeated remediation leaching on the groundwater level and mineralization should be conducted in the future, due to the possibility of accelerated salinization, fertility decline, and reduced yields of agricultural crops.

Keywords: irrigation; groundwater level; salinity; soil salinization; salt balance; Uzbekistan

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1 Introduction

Agriculture is an important economic sector in Uzbekistan, accounting for about 23% of the GDP and 27% of the total employment (Kulmatov, 2014). Irrigated farming produces more than 90% of the agricultural yield and consumes more than 90% of the nation's available water resources annually (Uzbekistan N C, 2009; Dyhovniy and Schutter, 2011; Opp et al., 2016). However, due to the flat topography of many irrigated areas of the Aral Sea Basin (ASB) countries, and the Republic of Uzbekistan in particular, which has a low level of natural drainage and extensive areas of flood irrigation with poor-quality water, soil salinization and degradation of the irrigated land has occurred (Panin, 1958; Pankova et al., 1996; Gafurova et al., 2005; SCNP, 2008; Kulmatov, 2018). Currently, more than 60% of all the irrigated areas in Uzbekistan are affected by salinization, which has resulted in a reduction in crop yields (e.g., 20%–30% for cotton). In most farms, the increased reclamation of leached water and application of large quantities of mineral fertilizers usually overcomes these losses (SCNP, 2008; Kulmatov, 2014, 2018). Agricultural, industrial and urban development have all contributed to the increased salinity in streams and rivers, but the likely effects of future development and climate change are unknown. Most of the change is related to the extent of human land uses, with climate change accounting for only 12% of the increase (Olson, 2019). However, land salinization and degradation are aggravated by the vulnerability of the sensitive arid lowlands to climate change.

Uzbekistan is located in an arid plain and the semi-arid foothill areas of the endorheic (closed) ASB. Being the main water consumer in the ASB, it receives about 80%–85% of its water resources from neighboring upstream countries, mainly through three transboundary rivers: the Syrdarya, the Amudarya, and the Zarafshan (FAO, 2017). For these and many other rivers of the ASB, water from snowmelt and glaciers in the Pamir and Tian Shan mountains is the main source of river discharge, and therefore their runoff is strongly dependent on changes in precipitation patterns and seasonal snow water storage capacity. The amount of precipitation received in recent years has decreased across the entire region, especially in the western parts of Turkmenistan, Uzbekistan and Kazakhstan (Bai et al., 2011; Lioubimtseva et al., 2018). Endorheic basins are known to be sensitive to various changes in ambient conditions, which makes them vulnerable to climate change and other human-induced pressures, including agricultural intensification (Karthé et al., 2017). Water abstraction for irrigation is the main cause of reduced river flows in downstream areas, but climate change also leads to a decrease in water discharges in river catchments. Temperature increases, followed by a spread in aridity have been observed during the last century throughout Central Asia, and this is predicted to continue in the future at a rate above the global average (Lioubimtseva and Henebry, 2009; Mannig et al., 2013). There is already clear evidence of the negative consequences of climate change, including a decrease in available water resources and a decline in agricultural productivity. Between the 1960s and the 2010s, the glaciated area decreased by 23%–49% in various river catchment basins of the ASB (Semakova et al., 2015). During the period of 2001–2010, the estimated volume of total glacier runoff in the Amudarya and Syrdarya river basins was reduced from 19.0 and 3.4 km³ to 18.0 and 3.2 km³, respectively (Savoskul and Smakhtin, 2013).

In the future, upstream countries are planning to construct facilities to accumulate additional water volumes in their existing water reservoirs for the production of hydropower energy, which may present another potential threat for land use sustainability. In combination with the high population growth rate in the ASB, this is likely to lead to a decrease in the availability of river water and an increase in water consumption (Uzbekistan N C, 2009; Dukhovny and Schutter, 2011; Groll et al., 2015). These trends are likely to be exacerbated until coordinated mitigation and adaptation activities are implemented (Dukhovny and Schutter, 2011; Groll et al., 2015; Canedo-Arguelles et al., 2016).

Furthermore, the extensive area of unsustainable land use and insufficient maintenance of the irrigation and drainage networks has caused an increase in the groundwater level and

mineralization. As a result, there has been a dramatic deterioration in the condition of irrigated areas in recent decades (SCNP, 2013; Kulmatov, 2014, 2018).

A comparative analysis of the climatic characteristics of desert and arid lands in Central Asia (Turan Depression, the Gobi Desert, and the deserts of the Dzungar and Tarim depressions) revealed that the degree of aridity, the degree of continentality, and the precipitation amount and regime differ among these areas (Pankova and Konyushkova, 2013). The deficit of moisture in the modern climate results in the preservation of salt accumulations at their point of origin. The importance of specific features of the climate in these regions has been noted, including the precipitation regime and the effect of the redistribution of salts in the profiles of automorphic salt-affected soils (Pankova and Konyushkova, 2013).

In recent years, there has been a deterioration in the quality and a decrease in the quantity of surface water resources and reclamation conditions in the irrigated lands of the provinces located in the mid- and downstream areas of the Amudarya River (Ibrakhimov et al., 2007). Only in Navoi Province in Uzbekistan between 2000 and 2015 (Kulmatov et al., 2018) has the groundwater level and soil salinity shown a slight improvement (Kulmatov et al., 2015).

In irrigated land in Jizzakh Province (Syrdarya River Basin) the amount of humus and the absorption capacity has declined in recent years, and at the same time the mechanical composition and fertility of soils has decreased (Sherimbetov, 2015; Rakhmatov and Abdullaev, 2016). Some studies have shown that long-term irrigation has a significant impact on the absorption capacity and physical and chemical properties of soils. When agricultural activities are conducted in irrigated soils for long periods, there are varying degrees of change in the soil layers.

It has been revealed that of the four soil components (humus, carbonates, gypsum, and water-soluble salts), water-soluble salts play the dominant role in spectral reflectance from saline soils (Karavanova et al., 2001; Sherimbetov, 2015).

Previous studies have attempted to investigate the relationships between soil productivity and salinization to ensure the sustainable use of water and land resources in the study area. Unfortunately, most of these studies were conducted within a short period of time, and the ionic composition and mineralization of the irrigation water over long periods was not sufficiently taken into account. The salt balance in the "irrigation water–soil–drainage water" system has not been studied. The pathways of salt migration in natural and agricultural land are important to understand for long-term monitoring and the prevention of soil and water salinization; however, they have still not been quantitatively determined in Jizzakh Province.

The objective of this study was to determine the long-term changes (1995–2016) in the quality of water used for irrigation, the level and mineralization of groundwater, the amelioration required in irrigated land, and the salt balance in the "irrigation water–soil–drainage water" system, as well as to develop recommendations for the sustainable use of water and land resources, with a specific focus on the Jizzakh irrigation zone.

2 Study area

Jizzakh Province is located in the center of the Republic of Uzbekistan. The province is bordered by Syrdarya Province in the east, the Republic of Kazakhstan in the north, and the Republic of Tajikistan in the south, and is enclosed by the Syrdarya River, the Turkestan border, and the Kizil-Kum Desert. The total area of irrigated land ($39^{\circ}57'–41^{\circ}24'N$, $66^{\circ}66'–68^{\circ}57'E$) in the province is 3.0×10^5 hm^2 and the main agricultural crops are: cotton, wheat, berries, fruit, grapes, melons and gourds.

The province is situated in a continental area that is characterized by dry and hot summers and moderately cold winters. Precipitation occurs mainly in winter and spring, with an average annual rainfall of 308 mm. The highest recorded air temperature was $36.4^{\circ}C$, which was observed in July. The average annual air temperature is $15.6^{\circ}C$. The humidity level is 70%–80% in winter. The vegetation growth period is 210–240 d (State Department of Statistics of Uzbekistan, 2017).

The soils of Jizzakh Province are classed as light gray meadow-brown, dark-brown, typical and light-brown grass, meadow-grass, grassland and swampy-grass, sandy desert, sand, and sandy-

saline soil (SCNP, 2016). The soils in the irrigation zone are classed as typical dark-brown and light-brown, meadow-grassy, grass-grassland, and meadow and wetland-grass soils.

The water supply in Jizzakh Province is transboundary and is largely dependent on the quantity of water coming from the countries containing the upper reaches of the main river sustems, i.e., Kyrgyzstan and Tajikistan. There are three main sources of irrigation water in the province with average annual volumes of diverted water of 1.89 km³ (62.2% of the total) from the Syrdarya River, 1.05 km³ (34.5%) from the Zarafshan River, and 0.10 km³ (3.3%) from the Sangzor River. The total water volume used in the province is 3.04 km³, of which 2.85 km³ (93%) is used for agricultural purposes and the remaining 0.19 km³ (7%) is used by industry and as the drinking water supply.

The organic matter concentration in water from the Syrdarya and Zarafshan rivers is very low; however, both rivers have elevated concentrations of inorganic substances, mainly in the form of sulfate, chloride, and carbonate ions (Table 1), which causes the secondary salinization of agricultural land (Yakubov et al., 2011; Karimov et al., 2014; Karimov et al., 2019).

Table 1 Mineralization of river water used for irrigation in Jizzakh Province during 2000–2016

River	Hydropost	River water mineralization (g/L)											
		2000			2005			2010			2016		
		Min	Max	Ave	Min	Max	Ave	Min	Max	Ave	Min	Max	Ave
Syrdarya	Nadejdinskiy	0.8	1.8	1.3	0.5	1.5	1.0	0.7	1.5	1.1	0.7	1.7	1.2
Zarafshan	Pervomayskiy	0.3	0.4	0.4	0.3	0.3	0.3	0.3	0.4	0.4	0.3	0.5	0.4

Note: Min, minimum; Max, maximum; Ave, average. Hydropost – measuring infrastructure within the main canal.

3 Materials and methods

The study used a database of field measurements of groundwater, mineralization and soil salinity levels conducted by the provincial Hydro-Geological Reclamation Expeditions (HGRE) under the Uzbek Basin Irrigation System Administration of the Ministry of Agriculture and Water Resources of the Republic of Uzbekistan. The HGRE conducts a monitoring program throughout Uzbekistan. Indicators of spatial and temporal changes in the levels and mineralization of groundwater in irrigated areas of Jizzakh Province were determined by HGRE specialists through the use of 2255 continuous monitoring wells. The Ministry of Agriculture and Water Resources also has a detailed database of long-term inorganic fertilizer use in all irrigated land in Uzbekistan, including Jizzakh Province, which has also been processed statistically and can be used to study the dynamics of fertilizer use. This existing extensive database was complemented by additional data regarding soil characteristics, soil salinity and cropping patterns based on the results of a soil-sampling exercise conducted by the authors.

Groundwater samples were taken by on-farm technicians, allowing the swift collection and analysis of more than 12,600 groundwater samples per year. The groundwater level monitoring stations were equipped with a tube (inner diameter of 90–110 mm, length of 3–6 m), filled with sand-gravel filters. Samples of groundwater from 0.0–1.0, 1.0–1.5, 1.5–2.0, 2.0–3.0, 3.0–5.0 and 10.0 m deep monitoring wells were taken and sent to the laboratory for analysis.

The determination of groundwater level and mineralization was performed three times a year on 1 April, 1 July and 1 October. The total mineralization of groundwater was determined using a portable electric conductometer (Progress 1T, Central Asian scientific research institute of irrigation (SANIIRI), Uzbekistan) and the chloride concentration was determined using the Mohr method, i.e., argentometric titration. During the first sampling period (April), the effects of extensive salt leaching, which occurred just before the irrigation season started, could be detected. The second sampling period (July) covered the period of peak irrigation activity and the third sampling period (October) took place immediately after the end of the growing season. This enabled an analysis of the lowering of the phreatic surface without groundwater recharge. An analysis of the groundwater level dynamics outside the growing season is important, because a seasonal salinity restoration might occur when the upward flux prevails over the lateral outflow. During 2000–2016, 111,634 undergroundwater samples were analyzed. Groundwater

mineralization was assessed and classified following the approach of Priklonsky (1970) (Table 2).

Table 2 Classification of groundwater (GW) based on total mineralization (Priklonsky, 1970)

Category	Total dissolved solids (TDS; g/L)
Fresh	0–1
Low mineralization	1–3
Medium mineralization	3–10
High mineralization	10–50

A total of 17,500 soil samples was taken each year by HGRE staff at the end of the growing season in November during 2000–2016. Each soil sample consisted of three subsamples, collected at depth of 0–30, 30–70 and 70–100 cm. More than 116,000 soil samples were collected from these depths in the irrigated areas to determine soil salinity. Each sample was considered to be representative of an area of 10–20 hm², resulting in a much denser grid of soil data than groundwater data.

Mineral fertilizers (mainly nitrogen-containing) can cause salinization of irrigated soils. In the calculation of the salt balance, the contribution of mineral fertilizers was included in the total balance of salts because their residues enter the collector-drainage network, and therefore their contribution was not taken into account separately. The effects of side flows, for example, unsaturated groundwater from irrigated land, were not taken into account when calculating the salt balance.

The soil salinity level of irrigated land was measured twice a year on 1 April and 1 October. For each soil sample, the electric conductivity was measured at the four corners of a 1.5×2.0 m² area surrounding the soil-sampling site. The soil samples were dried naturally in a location with no direct sunlight. The soil samples were then passed through a special sieve, with a mesh size of 0.1 cm, and three solution samples were prepared with 50 g of soil in distilled water and analyzed. The analyses of the soil salinity were conducted by applying two different methods: first through the extraction and assessment of the soluble salt content and then through an conductometer (IKS Express 1T, SANIIRI, Uzbekistan) applied to a 1:1 mixture of soil sample and water (Shirokova and Chernyshev, 1999).

The measured soil salinity (total mineralization) was then categorized as low (0.3–1.0 g/L), moderate (1.0–2.0 g/L), high (2.0–3.0 g/L), or very high (>3.0 g/L). This classification was then assigned to the area represented by each soil sample to obtain spatial and temporal information about the soil salinity distribution and dynamics. Based on the groundwater and soil data, the overall extent of the salinization of the irrigated areas was assessed using the salinity classification system developed by Bazilevich and Pankova (1970) (Table 3).

Table 3 Classification of soil salinity, based on TDS and chloride (Cl⁻) concentration (modified after Bazilevich and Pankova (1970))

No	Level of salinization	Sulfate	Chloride-sulfate		Sulfate-chloride		Chloride
		TDS (g/L)	TDS (g/L)	Cl ⁻ (g/L)	TDS (g/L)	Cl ⁻ (g/L)	Cl ⁻ (g/L)
1	No salinity	<0.3	<0.1	<0.01	<0.1	<0.01	<0.01
2	Low salinity	0.3–1.0	0.1–0.3	0.01–0.05	0.1–0.3	0.01–0.04	0.01–0.03
3	Moderate salinity	1.0–2.0	0.3–1.0	0.05–0.20	0.3–0.6	0.04–0.20	0.03–0.10
4	High salinity	2.0–3.0	1.0–2.0	0.20–0.30	0.6–1.0	0.20–0.30	0.10–0.20
5	Very high salinity	>3.0	>2.0	>0.30	>1.0	>0.30	>0.20

4 Results and discussion

4.1 Groundwater level and its influence on the characteristics of the irrigated land

In arid areas, factors affecting the efficiency and sustainability of irrigated agriculture are closely linked with the groundwater level. Improper irrigation and inefficient drainage networks can lead

to soil salinization (Eshchanov, 2008; Martin and Sauerborn, 2013; Kulmatov, 2018). To ensure sustainable long-term irrigation and crop yields, it is necessary to determine the reasons for the temporal and spatial changes in GW level and mineralization. A high groundwater salinity in combination with a shallow groundwater level leads to the salinization of irrigated land and to the bogging of the root system of crops, resulting in yield losses (Gafurova et al., 2005; Eshchanov, 2008).

Table 4 shows the admissible levels of groundwater salinity for different groundwater levels. A groundwater level of less than 1.0 m is especially critical because even salt concentrations of less than 1.0 g/L can lead to land salinization, while groundwater at depths of 3.0 m or more could have a salinity of up to 5.0 g/L before the full salinization effect sets in (Gafurova et al., 2005; Kulmatov et al., 2018).

These "critical depths" were used to assess the level of amelioration required for irrigated land and to conduct practical measures against the salinization of irrigated land. If groundwater was located above "the critical depth", the upward movement of salts from the lower layers would increase and the soils would be exposed to salinization. However, if the groundwater was located below the critical depth, the salt would not rise from the lower layers into the topsoil and soil salinization would not occur.

Table 4 The critical groundwater (GW) depth and the related admissible level of groundwater salinity in irrigated soils (Gafurova et al., 2005)

Critical GW depth (m)	Admissible level of GW salinity (g/L)	Maximum concentration of chlorine in GW (%)
0.8–1.0	About 1.0	0.17
1.0–1.5	1.0–2.0	0.17–0.27
1.5–2.5	2.0–3.0	0.27–0.37
2.5–3.5	3.0–5.0	0.37–0.69
>3.0	>5.0	>0.69

The critical groundwater depth also depends on the soil properties, especially the capillary structure, the water retention potential, and the percolation characteristics (e.g., the amount of macropores or the soil density) (Childs, 1969; Nielsen et al., 1973; Bowles, 1979; Heath, 2004; Gafurova et al., 2005; Kulmatov, 2018). To take these parameters into account, Table 5 shows the critical groundwater depth in relation to the water retention and percolation capability of the irrigated land. This integration of several important soil hydrological parameters makes the critical groundwater depth a crucial concept for the sustainable use of the irrigated land resources.

Table 5 Critical GW depth related to the water retention and percolation capability of irrigated soils (Gafurova et al., 2005)

Water retention and percolation capability of irrigated soils	GW salinity (g/L)	Critical GW depth (m)	Admissible GW level (m)
Weak	1.5–3.0	1.2–1.5	1.5–1.8
Weak	3.5	1.5–1.8	1.8–2.0
Average	3.5	1.8–2.0	2.0–2.3
Strong	3.5	2.0–2.2	2.3–2.5
Strong	3.0–8.0	2.2–2.5	2.5–3.0

To ensure long-term sustainable irrigation in farming and stable crop yields, it is necessary to determine the reasons for temporary changes in the groundwater level, with special consideration given to the critical groundwater level and salinity. An analysis of the dynamics of groundwater level changes in Jizzakh Province (Fig. 1) revealed that there were no irrigated fields with very shallow (0.0–1.0 m) groundwater levels. The largest area of all the irrigated land in the province ($200.0 \times 10^3 \text{ hm}^2$; 66.96% of the total irrigated land) had a groundwater table at a depth of 2.1–3.0 m; followed by an area of $63.93 \times 10^3 \text{ hm}^2$ (21.41%) at 3.1–5.0 m; $2.71 \times 10^3 \text{ hm}^2$ (0.927%) at 5.0 m; $6.23 \times 10^3 \text{ hm}^2$ (2.09%) at 1.5–2.0 m; and $0.80 \times 10^3 \text{ hm}^2$ (0.27%) at 1.0–1.5 m. During 1995–2016

the groundwater level in a large area of the province's irrigated land ($291.00 \times 10^3 \text{ hm}^2$) remained generally stable, varying only between 5%–10%.

The groundwater level in Jizzakh Province has also been elevated by anthropogenic activities, which have led to additional inputs of water into the groundwater zone. Extensive leaching during spring, high-intensity irrigation during the vegetative season, leakage from irrigation canals and drainage water collectors, and waterlogging due to damaged or blocked drainage networks are the main reasons for the raised groundwater levels.

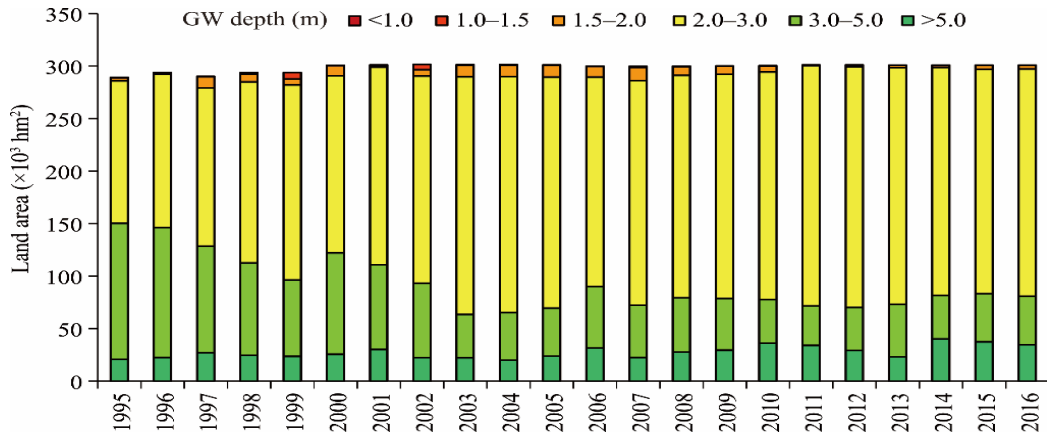


Fig. 1 Dynamics of groundwater (GW) level changes in irrigated land in Jizzakh Province

This can be problematic for the growth of crops, because shallow saline groundwater levels can lead to reduced water uptake (e.g., in cotton), and thus to crop losses (Hutmacher et al., 1996; SANIIRI, 2005). Studies conducted by the Central Asian Research Institute of Irrigation (SANIIRI) have shown that the productivity of cotton decreases by 15%–20% in slightly saline soils, by 35%–40% in moderately saline soils, and by 70%–80% in highly saline soils. In addition to the reduced yields (from 2.10 to 0.90–1.25 t/hm²), the salinity has also been shown to influence the quality of the raw cotton fibers (SANIIRI, 2005).

4.2 Dynamics of groundwater mineralization fluctuations in the province

The groundwater observations for 1995–2016 indicated that the average groundwater mineralization in most of the irrigated land of the province fluctuated between 1.1–5.0 g/L (75.3% of all irrigated land), with an average of less than 1.0 g/L in 13.1% of the irrigated land and a range of 5.1–10.0 g/L in 10.5% of the irrigated land (Table 6).

Table 6 Dynamics of the long-term (1995–2016) average GW mineralization in Jizzakh Province

Mineralization (g/L)	Irrigated land area ($\times 10^3 \text{ hm}^2$)	Percentage (%)
≤ 1.0	39.17	13.1
1.1–3.0	116.15	38.9
3.1–5.0	108.85	36.4
5.1–10.0	31.48	10.5
≥ 10.0	3.01	1.0
Total	296.88	100.0

In 1995, the irrigated area with a groundwater mineralization of 0.0–3.0 g/L was $73.65 \times 10^3 \text{ hm}^2$, while in 2016 the corresponding figure was $152.76 \times 10^3 \text{ hm}^2$, i.e., an increase of $79.11 \times 10^3 \text{ hm}^2$ (25.30% of all irrigated land). The area with an average mineralization (3.1–5.0 g/L) also increased from $86.11 \times 10^3 \text{ hm}^2$ in 1995 to $133.33 \times 10^3 \text{ hm}^2$ in 2016, i.e., an increase of $47.22 \times 10^3 \text{ hm}^2$ (14.57%).

In contrast, in 1995 the irrigated area with a high groundwater mineralization (5.0–10.0 g/L) was $118.93 \times 10^3 \text{ hm}^2$, while in 2016 it was $13.75 \times 10^3 \text{ hm}^2$, i.e., a considerable decrease of 105.18×10^3

hm² (36.58%). Similarly, the irrigated area with a groundwater mineralization of 10.0 g/L and above was reduced from 10.32×10^3 hm² in 1995 to 0.70×10^3 hm² in 2016 (97.00%, Fig. 2). In most of the irrigated land area of the province (75.33%) the groundwater mineralization was in the range of 1.1 to 3.0–5.0 g/L.

Within the period 1995–2016 the area with a relatively low level of groundwater mineralization increased from 58.70×10^3 to 91.00×10^3 hm², which indicates a slight improvement in the state of the irrigated land of the province.

During the same period, the area of irrigated land with an average groundwater mineralization (3.1–5.0 g/L) decreased from 44.90×10^3 to 16.50×10^3 hm² (20.80%). This indicates that there were positive changes in the dynamics of groundwater mineralization in the irrigated areas of the province. In 1995, the area with low groundwater mineralization (0.0–3.0 g/L) was 73.65×10^3 hm², while in 2016 it was 152.76×10^3 hm², i.e., an increase of 79.11×10^3 hm² (25.34%). The area with an average groundwater mineralization of 3.1–5.0 g/L was 86.11×10^3 hm², while in 1995 it was 133.33×10^3 hm² in 2016, i.e., an increase of 47.22×10^3 hm² (14.57%). The irrigated area with a high groundwater mineralization (5.0–10 g/L) decreased from 118.93×10^3 to 13.75×10^3 hm² during 1995–2016, and therefore the GW quality was improved over 105.18×10^3 hm² (36.58%). The area of irrigated land with an average groundwater mineralization of 10.0 g/L or more decreased from 10.32×10^3 hm² in 1995 to 0.70×10^3 hm² in 2016 (i.e., a decrease of 9.62×10^3 hm² or 97.00%). The area with a high groundwater mineralization (5.0–10.0 g/L) was 118.93×10^3 hm² in 1995, but this decreased to 13.75×10^3 hm² in 2016, i.e., groundwater quality was improved over an area of 105.18×10^3 hm² (63.00%).

Thus, based on the results of these analyses it can be concluded that the quality of groundwater in Jizzakh Province has considerably improved in recent years. One of the main causes of the wedging of the groundwater outflow in the province is the improvement of the drainage network in the Jizzakh irrigation area, which has diverted drainage water toward the Aydar-Arnasay and Tuzkan lakes in the Mirzachul Desert in the southern part of the province.

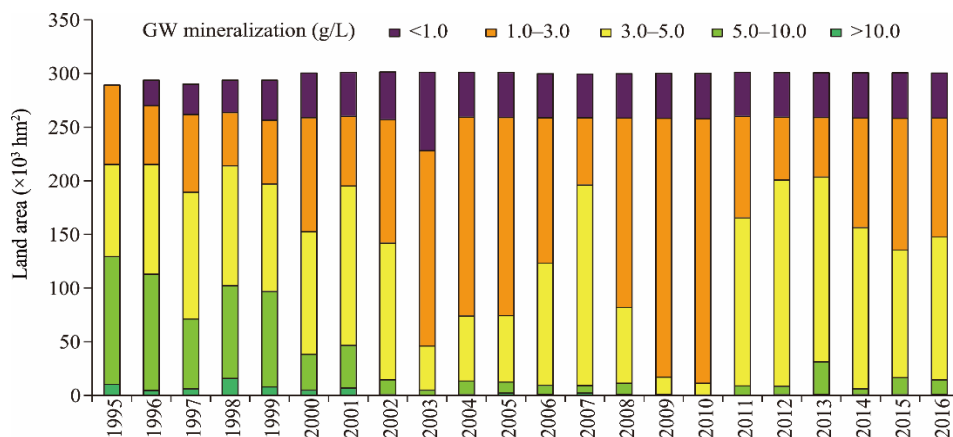


Fig. 2 Long-term (1995–2016) dynamics of groundwater (GW) mineralization of the irrigated land in Jizzakh Province

4.3 Salinization level of irrigated soils in the province

As shown in Figure 1, the majority of the irrigated area in Jizzakh Province was characterized by a groundwater level of more than 2.0 m. In addition, the majority of the irrigated area in the province was characterized by a low groundwater mineralization (Fig. 2).

However, because the availability of low mineralized deep groundwater is limited, the additional water needed for crop growth in the irrigated areas of the province is provided by water diversion from the Syrdarya and the Zarafshan rivers. The level of mineralization in the water of both these rivers has increased (Table 2), with the result that the extensive irrigation has led not only to a rising groundwater level but also an increase in groundwater mineralization.

An efficient drainage network would prevent salt accumulation, but the technical infrastructure

used in the drainage system in Jizzakh Province is outdated and not always well maintained; therefore, salt removal is incomplete.

The results of this study indicated that most of the irrigated land in Jizzakh Province is slightly saline. On average 51.3% of the irrigated area fell into this category (Fig. 3). Another 17.7% showed no sign of salinization. However, 29.0% of the irrigated area was found to be moderately saline and 2.0% was highly saline. Overall, salinization in Jizzakh Province was much lower than in neighboring Bukhara Province, where 27.6% of the irrigated area is classified as moderately saline and 7.8% as highly saline (Kulmatov et al., 2015).

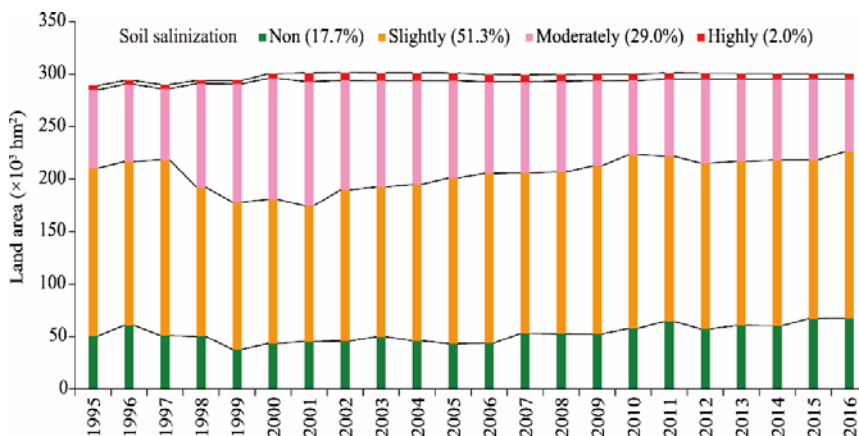


Fig. 3 Temporal soil salinization dynamics of the irrigated area in Jizzakh Province from 1995 to 2016

All the relevant parameters (irrigation water mineralization, groundwater level, and groundwater salinity) are subject to considerable temporal dynamics (Table 2; Figs. 1 and 2), which means that their influence on the soil salinity is also dynamic. Therefore, managing the salinization of irrigated land is a complex process and it is necessary to conduct flexible agrotechnical practices. In recent years, this paradigm has been increasingly incorporated into the agrarian sector, explaining the observed improvements in the salinity conditions in Jizzakh Province.

Salinization of agricultural land negatively affects the soil structure and biological activity of soil microorganisms. This can lead to a decline in the use of mineral fertilizers by plants and may cause various plant diseases (SCNP, 2013). Cotton productivity in Uzbekistan has decreased by 15%–20% in areas with low-salinity soils, 35%–40% in areas with moderate salinity soils, and 70%–80% in areas with high-salinity soils (SANIIRI, 2005). Therefore, monitoring and assessment of the salinization level of irrigated soils is very important for agriculture. Based on the results obtained, programs to control the effects of salt leaching on the following year's crop can be initiated.

In Jizzakh Province in 1995 the area of irrigated land not affected by salinization was 50.57×10^3 hm^2 (17.50% of the total irrigated land area), which increased to 67.64×10^3 hm^2 (22.50%; Fig. 3) in 2016. This indicates that in recent years, the irrigated area that is not affected by salinization increased by 17.07×10^3 hm^2 (5.00%). The area of low-salinity irrigated land in 1995 was 159.81×10^3 hm^2 (55.30%), which decreased to 158.67×10^3 hm^2 (52.80%) in 2016, i.e., a slight decrease of 1.15×10^3 hm^2 (0.50%). The area of moderate-salinity irrigated land in 1995 was 74.59×10^3 hm^2 (25.80%), which decreased to 68.62×10^3 hm^2 (22.80%) in 2016, i.e., a decrease of 5.97×10^3 hm^2 (3.00%). The high-salinity irrigated land area slightly increased from 4.04×10^3 hm^2 (1.40%) in 1995 to 5.62×10^3 hm^2 (2.00%) in 2016.

In general, it can be concluded that during the period of 1995–2016 the salinization level of irrigated lands in the province have slightly increased. Based on the average soil salinization level, the province's irrigated areas could be classified as follows: no salinity (17.7%), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%) (Fig. 3). The observed slight increase of salinity in irrigated land in Jizzakh Province occurred due to the use of outdated irrigation technology and inefficient drainage systems. Because of the relatively high amounts of

mineralized salts in irrigated land and the proximity of groundwater, reclamation leaching of the irrigated land is required in winter each year.

4.4 "Irrigation water–soil–drainage water" salt balance

Climate change and the frequently recurring droughts in Central Asia in recent years have diminished the volume of river water flows, with the result that more water resources are required for irrigation (SCNP, 2013). If the amount of water available for irrigation does not meet the scientifically based needs of the crop, or if excessive amounts of water are used, salts accumulate in irrigated soils. The accumulation of salts associated with irrigated agriculture is mainly caused by two factors. One is the introduction of salts in the irrigation water to the land. The second is the extremely high level of saline groundwater, caused by excessive irrigation and a poor drainage system (Dukhovny, 1983; Kovda, 2008). It is difficult for excess water to flow-out naturally from irrigated areas, which leads to salinization and swamping, and reduces soil fertility (Kulmatov et al., 2015, 2018). Increased salinity in water and soil has had serious negative impacts on agriculture and food security, as well as human health, all over the world. The direct impact of salinity on both the soil microbial community and crops can have a serious impact on crop production, yield, and cropping pattern (Yan et al., 2015; Rahaman et al., 2019).

It is therefore important to identify the total amount of salts entering the irrigated land of the province with irrigation water, the amount of salt remaining in the irrigated areas, and the amount of salt leached out through the drainage system (reclamation leaching). An analysis of the dynamics of the "irrigation water–soil–drainage water" salt balance in the irrigated area of the province was conducted based on the results of the monitoring data for 2000–2016 (Table 7).

During the period of 2000–2016, an average 2690.9×10^6 m³ of water was used for irrigation during the vegetation growth period, autumn crop irrigation, and winter salt leaching in the province. The highest volume of irrigation water (3368.8×10^6 m³) was used in 2005, from which 799.0×10^6 m³ of water was diverted to the drainage systems (Table 7).

The average annual amount of water used for irrigation (2690.9×10^6 m³) had a total salt content of 3271.4×10^3 t, with chlorides accounting for 974.9×10^3 t. The annual average amount of water flowing out as drainage water was 826.4×10^6 m³, which contained a total salt content of 2878.0×10^3 t, with chlorides accounting for 887.5×10^3 t. Thus, the total amount of salts that remained in the irrigated area was 393.4×10^3 t, with chlorides accounting for 87.4×10^3 t.

On average for each hectare of irrigated land in the province 1.21 t of salts entered the soil with irrigation water and 1.06 t of salts were leached out by drainage water. Thus, 12.0% of the salts and 8.9% of the chlorides that entered the soil with irrigation water remained in the irrigated areas of the province. On average, 0.15 t of salt remained on each hectare of irrigated land (Table 7). The lowest amount of irrigation water used was 2351.8×10^6 m³ in 2001 (an extreme drought year in the country and the province), 980.6×10^6 m³ of which was diverted to the drainage system. The remaining 1371.2×10^6 m³ of water was used in the vegetation growth season. In autumn–winter, 88.9×10^6 m³ of river water is used each year for reclamation leaching in the irrigated soils in the province. This has resulted in desalinization and an improvement in the reclamation status of the irrigated land in the province.

The average salt concentration in water used in the province for irrigation purposes was 1.2 g/L in 2000–2016. The maximum amount of salts (4099.2×10^3 t) that entered the soils of the irrigated area occurred in 2000, of which 4319.1×10^3 t was leached out by drainage water. Thus, the amount of salt entering the irrigated area in irrigation water was less than the amount of salt leached out from the irrigated area by drainage water. This phenomenon likely occurred because salt leaching in the irrigated areas in the autumn–winter period was achieved more effectively in 2000 than in other years.

In 2012, the total amount of salt entering the soil through irrigation water was 2865.8×10^3 t, of which 2209.1×10^3 t was leached out by drainage water. Thus, only 656.7×10^3 t of the salts remained in the irrigated area. In 2000–2016, the average salt content in the irrigation water was 3271.4×10^3 t, of which 2878.0×10^3 t was leached out from the irrigated land. The total amount of residual salt in the irrigated area was 393.4×10^3 t. As a result, 1.3 t/hm² salt of irrigated land has

accumulated during the 17 years studied. Thus, 11.7% of the total incoming salts have remained in the irrigated area (Table 7).

Table 7 "Irrigation water–soil–drainage water" salt balance of Jizzakh Province during 2000–2016

Year	Irrigation water input					Drainage water output					Salt balance	
	WV	M	C	TS	TC	WV	M	C	TS	TC	TSB	TCB
	($\times 10^6$ m ³)	(g/L)		($\times 10^3$ t)		($\times 10^6$ m ³)	(g/L)		($\times 10^3$ t)		($\times 10^3$ t)	
2000	2644.6	1.5	0.4	4099.2	925.6	966.5	4.5	1.1	4319.1	1111.4	−219.9	−185.8
2001	2351.8	1.3	0.4	3125.5	945.4	980.6	4.2	1.0	4166.5	950.1	−1040.9	−4.7
2002	2507.8	1.3	0.4	3265.1	942.9	935.4	3.4	1.1	3177.7	1028.9	87.4	−86.1
2003	2600.8	1.1	0.3	3081.9	847.9	830.8	3.3	0.9	2789.9	717.0	292.0	130.9
2004	3067.4	1.1	0.3	3542.8	1015.3	920.5	4.6	1.2	4207.7	1150.7	−664.9	−135.4
2005	3368.8	1.0	0.3	3665.3	1152.1	798.9	2.3	0.8	1872.8	659.9	1792.5	492.2
2006	2660.4	1.1	0.3	3075.5	824.7	719.5	2.1	0.8	1548.5	603.7	1526.9	221.0
2007	2816.9	1.1	0.3	3188.8	997.2	765.4	3.1	0.9	2386.6	682.8	802.1	314.4
2008	2447.3	1.2	0.4	3054.2	1003.4	621.4	3.2	0.9	2015.8	587.2	1038.7	416.2
2009	2513.8	1.1	0.3	2928.5	852.2	692.7	2.9	0.8	2063.6	594.3	864.9	257.8
2010	2812.8	1.1	0.3	3220.6	992.9	732.6	3.6	0.9	2677.6	697.4	542.9	295.4
2011	2593.9	1.1	0.4	3001.1	1027.2	679.4	3.1	0.9	2081.6	607.4	919.4	419.8
2012	2390.2	1.2	0.4	2865.9	948.9	737.3	3.0	2.9	2209.1	2169.3	656.7	−1220.4
2013	3100.8	1.3	0.4	4046.6	1364.3	893.6	3.5	1.0	3120.6	895.4	925.9	468.9
2014	2548.8	1.1	0.3	2923.4	823.2	781.6	3.9	1.1	3040.3	886.3	−116.8	−63.0
2015	2464.5	1.2	0.3	2883.4	845.3	1115.8	3.7	0.9	4108.9	1102.4	−1225.5	−257.1
2016	2854.7	1.3	0.4	3645.4	1064.8	875.9	3.6	0.7	3140.2	643.8	505.3	421.0
Average	2690.9	1.2	0.3	3271.4	974.9	826.4	3.4	1.1	2878.0	887.5	393.4	87.4

Note: WV, water volume; M, mineralization; C, chlorine content; TS, total salt content; TC, total chlorine content; TSB, total salt balance; TCB, total chlorine balance.

During the period of 2000–2016, the average chlorine concentration in the water used for irrigation in the province was 0.3 g/L. The lowest amount of chlorine that entered irrigated fields (823.2×10^3 t) was observed in 2014, with 886.3×10^3 t of chlorine leached out through drainage waters. This implies that the amount of chlorine in the water used for irrigation was less than the amount flowing out in the drainage water. The likely reason for this was that soil leaching was more effectively conducted during the autumn–winter season in 2014 than in other years.

In 2013, the total amount of chlorine entering the soils in irrigation water was 1364.3×10^3 t, of which 895.4×10^3 t was leached out through drainage water. As a result, 468.9×10^3 t of excess chlorine remained in irrigated areas.

In general, during 2000–2016, the average amount of chlorides in water used for irrigation was 974.9×10^3 t/a, of which 887.5×10^3 t/a was leached out through drainage water. The amount of residual chlorides was 87.4×10^3 t/a, with 0.29 t/a of chlorides per hectare of irrigated land accumulating during the 17-a study period. On average, 10.1% of the total incoming chlorides accumulated in irrigated land (Table 7). Thus, the analysis of the "irrigation water–soil–drainage water" salt balance indicated a low salinization level in the irrigated soils of the region.

Most of the salt-affected land has experienced salinization due to natural causes, with the accumulation of salts over long periods of time in arid and semiarid zones (Rengasamy, 2002). More than 800×10^6 hm² of land throughout the world are salt affected, including 4.5×10^7 hm² (20%) of the current 2.3×10^8 hm² of irrigated land (FAO, 2008). We found that the average value

of groundwater salinization in most (75.3%) of the irrigated land of Jizzakh Province fluctuated between 1.1–5.0 g/L, with values less than 1.0 g/L in 13.1% of the irrigated land and in the range of 5.1–10.0 g/L in 10.5% of the irrigated land. A high groundwater salinization level can lead to an increase in the salt content of the upper layers of the soil through capillary action, which affects the root zone of crops, reducing their productivity. Salinization can be effectively controlled in irrigated land by reducing the groundwater level, increasing the efficiency of drainage networks that divert drainage water, and, most importantly, using modern climate-smart water-saving irrigation techniques.

Low rainfall in dryland areas, high transpiration rates by vegetation, and high evaporation rates during summer cause an accumulation of salts in the root zone layers (Rengasamy, 2002). Temperatures in the ASB are likely to increase by 2°C–3°C by 2050 and 3°C–5°C by 2080, and are predicted to be particularly high in summer and autumn (Lioubimtseva, 2015). Under an elevated soil salinity, repeated remediation leaching on irrigated land reduces the surface soil salinity, but because the existing drainage systems do not cover all of the irrigated land area and are inefficient, permanent salinization, a decline in fertility, and reduced yields of agricultural crops can occur. During the period of 1995–2016, the salinization level of irrigated lands in Jizzakh Province increased slightly and the area could be divided into the following classes: no salinity (17.7% of the total area), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%). Remediation leaching of soils with elevated salinity requires the use of additional water resources, which increases the groundwater level and mineralization, and also accelerates the salinization process in irrigated soils.

5 Conclusions

Under the current conditions of low soil fertility, stable crop yields (e.g., 10%–15% for cotton and grain) in Jizzakh Province have been achieved mainly due to the use of large volumes of mineral fertilizers. However, the permanent use of excess amounts of mineral fertilizers (primarily nitrogen-containing) can also cause the salinization of soils.

The average value of groundwater salinization in most (75.3%) of the irrigated land fluctuated between 1.1–5.0 g/L, with values less than 1.0 g/L in 13.1% of the irrigated land and in the range of 5.1–10.0 g/L in 10.5% of the irrigated land. The condition of the irrigated land in the province has not improved significantly. A high groundwater salinization level can lead to a reduction of crop productivity. The best way to effectively control salinization is to reduce the groundwater level, increase the efficiency of drainage networks, and, most importantly, use modern climate-smart irrigation techniques.

The salt balance in the "irrigation water–soil–drainage water" system is extremely complex. It changes over time and requires the monitoring and analysis of a large number of parameters. Repeated remediation leaching reduces surface soil salinity, but because the existing drainage systems do not fully cover the irrigated land area and are ineffective, permanent salinization, a decline in fertility, and reduced yields of agricultural crops are unavoidable. During the period of 1995–2016, the salinization level of irrigated land in Jizzakh Province has slightly increased and the area could be divided into the following classes: no salinity (17.7% of the total area), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%).

The analysis of the long-term monitoring data for Jizzakh Province revealed the importance of the sustainable use of the available water resources, as well as the need to improve the reclamation status of irrigated land, maintain the groundwater level and mineralization at an optimal level, keep irrigated soil salinization at an optimal level, and maintain the drainage system to ensure its effective operation. This would enable the sustainable use of water and land resources under a changing climate. Detailed studies of the salt balance in irrigated areas, and the impact of climate change, increased fertilizer use, and repeated remediation leaching on groundwater level and mineralization, should be conducted in the future to better understand the processes leading to accelerated salinization, declines in fertility, and reduced yields of agricultural crops.

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References

- Bai J, Chen X, Li J, et al. 2011. Changes in the area of inland lakes in arid regions of Central Asia during the past 30 years. *Environmental Monitoring and Assessment*, 178(1–4): 247–256.
- Bazilevich N I, Pankova E I. 1970. Classification of the Steppe by the Mineralization of Ground Waters. Moscow: Kolos Press, 196. (in Russian)
- Canedo-Arguelles M, Hawkins C P, Kefford B J, et al. 2016. Saving freshwater from salts. *Science*, 351(6276): 914–916.
- Dukhovny V A. 1983. Irrigation Complexes on New Lands of Central Asia. Tashkent: Uzbekistan Press, 184. (In Russian)
- Dukhovny V A, Schutter J L. 2011. Water in Central Asia: Past, Present, Future. New York: CRC Press, 432.
- Eshchanov R. 2008. The basis of agroecological and sustainable use of land and water resources (on the example of the Khorezm province). PhD Dissertation. Tashkent: Institute for agrochemistry and soil science of Uzbekistan Academy of Sciences. (in Russian)
- FAO (Food and Agriculture Organization of the United Nations). 2008. FAO Land and Plant Nutrition Management Service. [2019–08–11]. <http://www.fao.org/ag/agl/agll/spush>.
- FAO. 2017. Drought characteristics and management in Central Asia and Turkey. FAO Water Reports. Rome, Italy, 114.
- Gafurova L A, Abdullaev S A, Nomozov H K. 2005. Encyclopedia of Uzbekistan. Tashkent: National Publishing House, 190. (in Russian)
- Groll M, Kulmatov R, Mullabaev N, et al. 2016. Rise and decline of the fishery industry in the Aydarkul-Arnasay lake system (Uzbekistan): effects of reservoir management, irrigation farming and climate change on an unstable ecosystem. *Environmental Earth Sciences*, 75(10): 921.
- Groll M, Opp C, Kulmatov R, et al. 2015. Water quality, potential conflicts and solutions—an upstream–downstream analysis of the transnational Zarafshan River (Tajikistan, Uzbekistan). *Environmental Earth Sciences*, 73: 743–763.
- Ibrakhimov M, Khamzina A, Forkutsa I, et al. 2007. Groundwater table and salinity: Spatial and temporal distribution and influence on soil salinization in Khorezm region (Uzbekistan, Aral Sea Basin). *Irrigation and Drainage Systems*. 21(3): 219–236.
- Karavanova E I, Shrestha D P, Orlov D S. 2001. Application of remote sensing techniques for the study of soil salinity in semi-arid Uzbekistan. In: Bridges E M. Response to Land Degradation. New York: CRC Press, 261–273.
- Karimov B K, Matthies M, Kamilov B G. 2014. Unconventional water resources of agricultural origin and their re-utilization potential for development of desert land aquaculture in the Aral Sea basin. In: Bhaduri A, Bogardi J, Leentvaar J, et al. The Global Water System in the Anthropocene. Switzerland: Springer, Cham, 143–159.
- Karimov B K, Matthies M, Talskikh V, et al. 2019. Salinization of river waters and suitability of electric conductivity value for saving freshwater from salts in Aral Sea Basin. *Asian Journal of Water, Environment and Pollution*. 16(3): 109–114.
- Karthe D, Abdullaev I, Boldgiv B, et al. 2017. Water in Central Asia: an integrated assessment for science-based management. *Environmental Earth Sciences*, 76: 690, doi: 10.1007/s12665-017-6994-x.
- Kovda V A. 2008. Problems of Desertification and Salinization of Soils in Arid Regions of the World. Moscow: Nauka Press, 415. (in Russian)
- Kulmatov R. 2014. Problems of sustainable use and management of water and land resources in Uzbekistan. *Journal of Water Resource and Protection*, 6(1): 35–42.
- Kulmatov R. 2018. Sustainable Development Indicators of Lower Zarafshon Province and Their Practical Evaluation (Uzbekistan). Germany: LAP LAMBERT Academic Publishing, 110.
- Kulmatov R, Rasulov A, Kulmatova D, et al. 2015. The modern problems of sustainable use and management of irrigated lands on the example of the Bukhara province (Uzbekistan). *Journal of Water Resource and Protection*, 7: 956–971.
- Kulmatov R, Groll M, Rasulov A, et al. 2018. Status quo and present challenges of the sustainable use and management of water and land resources in Central Asian irrigation zones—The example of the Navoi region (Uzbekistan). *Quaternary International*, 464: 396–410.
- Lioubimtseva E, Henebry G M. 2009. Climate and environmental change in arid Central Asia: Impacts, vulnerability, and adaptations. *Journal of Arid Environments*, 73(11): 963–977.
- Lioubimtseva E. 2015. A multi-scale assessment of human vulnerability to climate change in the Aral Sea Basin. *Environmental Earth Sciences*, 73(2): 719–729.

- Mannig B, Müller M, Starke E, et al. 2013. Dynamical downscaling of climate change in Central Asia. *Global and Planetary Change*, 110: 26–39.
- Olson J R. 2019. Predicting combined effects of land use and climate change on river and stream salinity. *Philosophical Transactions of the Royal Society B*, 374(1764): 20180005, doi: 10.1098/rstb.2018.0005.
- Opp C, Groll M, Aslanov I, et al. 2016. Aeolian dust deposition in the southern Aral Sea province (Uzbekistan): Ground-based monitoring results from the LUCA project. *Quaternary International*, 429: 86–99.
- Panin P S. 1958. Processes of a Salt-out in the Washed-out Thicknesses of Soils. Novosibirsk: Siberian Branch of Nauka Press, 303. (in Russian)
- Pankova E, Aydarov J, Yamnova J et al. 1996. Natural and Anthropogenic Salinization of Soils in the Aral Sea Basin (Geography, Genesis, Evolution). Moscow: Soil Science Institute Named after Dokuchaev V V. 180. (in Russian)
- Pankova E I, Konyushkova M V. 2013. Climate and soil salinity in the deserts of Central Asia. *Eurasian Soil Science*, 46(7): 721–727.
- Ponomareva V V, Plotnikova T A. 1980. Humus and soil formation (methods and results of the study). Leningrad: Nauka Press, 222. (in Russian)
- Prikonsky V. 1970. The Methodical Recommendations on Mineralization of Solonetz Lands and Accounting Saline Soils. Moscow: Kolos Press, 185. (in Russian)
- Rahaman M A, Mohammad M R, Nazimuzzaman M. 2019. Impact of Salinity on Infectious Disease Outbreaks: Experiences from the Global Coastal Region. In: Leal Filho W, Wall T, Azul A, et al. *Good Health and Well-Being. Encyclopedia of the UN Sustainable Development Goals*. Springer, Cham. [2019-05-25]. https://doi.org/10.1007/978-3-319-69627-0_106-1.
- Rakhmatov Z, Abdullaev C. 2016. Changes in the physical and chemical properties of Jizzakh desert soils due to irrigation. *Bulletin of the Agrarian Science of Uzbekistan*, 2(64): 47–52. (in Russian)
- Rengasamy P. 2002. Transient salinity and subsoil constraints to dryland farming in Australian sodic soils: an overview. *Australian Journal of Experimental Agriculture*, 42(3): 351–361.
- SANIIRI (Central Asian Research Institute of Irrigation). 2005. Improvement of Monitoring of Salt Processes on the Irrigated Lands on the basis of Use of Modern Technologies and Development of Ways of Prevention of Damages to a Harvest of Crops from Salinity of Soils. NIR Report, 111. (in Russian)
- Savoskul O S, Smakhtin V. 2013. Glacier systems and seasonal snow cover in six major Asian river basins: hydrological role under changing climate. In: IWMI research report 150. Colombo, Sri-Lanka, 53.
- Semakova E K, Gunasekara Z, Semakov D. 2015. Identification of the glaciers and mountain naturally dammed lakes in the Pskem, the Kashkadarya and the Surhandarya River basins, Uzbekistan, using ALOS satellite data. *Geomatics, Natural Hazards and Risk*, 7(3): 1081–1098.
- Sherimbetov B. 2015. The results of research on the study of the soil-ecological state of the Jizzakh steppe based on GIS technology. *Bulletin of the Agrarian Science of Uzbekistan*, 4(62): 24–30. (in Russian)
- Shirokova Y I, Chernyshev A K. 1999. Rapid method for determination of soil salinity and water in conditions of Uzbekistan. *Journal of Agriculture of Uzbekistan*, 5: 45–52. (in Russian)
- SCNP (State Committee on Nature Protection). 2008. National Report on the State of the Environment and the Use of Natural Resources in the Republic of Uzbekistan (retrospective analysis for 1988–2008). Tashkent: Chinor ENK Press, 251. (in Russian)
- SCNP. 2013. National Report on the Conditions of the Environment and the Use of Natural Resources in the Republic of Uzbekistan (2008–2011). Tashkent: Chinor ENK Press, 255. (in Russian)
- SCNP. 2016. National Report on the state of land resources in the Republic of Uzbekistan, Tashkent, Uzbekistan: Chinor ENK Press, 73. (in Russian)
- State Department of Statistics of Uzbekistan. 2017. Annual Statistics 2000–2017 of the State Department of Statistics of Uzbekistan. [2017–12–16]. www.statistics.uz (in Russian).
- Uzbekistan N C. 2009. Second National Communication of the Republic of Uzbekistan under the United Nations Framework Convention on Climate Change. Tashkent, 189.
- Yakubov X E, Yakubov M A, Yakubov S X. 2011. Collector-Drainage Waters in Central Asia and Estimation of Their Use for Irrigation. Tashkent: IWP UzAS SIC ICWC Press, 189. (in Russian)
- Yan N, Marschner P, Cao W, et al. 2015. Influence of salinity and water content on soil microorganisms. *International Soil and Water Conservation Research*, 3(4): 316–323.